PHOTOSYNTHETICALLY AVAILABLE RADIATION (PAR) IN THE COLORADO RIVER: GLEN AND GRAND CANYON

DRAFT REPORT

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TABLE OF CONTENTS

List of Tables	ii
List of Figures	iii
Abstract	1
Introduction Background Information Problem Statement Objectives	2 2 3 4
Materials and Methods Study Area Equipment and Data Collection Data Analysis	5 5 9
Analysis of Vertical Light Attenuation Sediment Concentration Effects on Attenuation of PAR Distance Effects on Attenuation of PAR Light Absorbance and Scattering Optical Light Conditions within the Euphotic Zone	12 12 12 16 19 22
Discussion Sediment Concentration Effects on Attenuation of PAR Distance Effects on Attenuation of PAR Light Absorptance and Scattering Optical Conditions within the Euphotic Zone Glen Canyon Dam Operations	27 28 29 31 34 37
Conclusions	38
Literature Cited	40
Acknowledgements	44
Appendix A	45 46
Appendix B	52 53 59
Appendix C	61 62
Appendix D	67 68

LIST OF TABLES

- Table 1 Analysis on integrated sediment concentrations (g/L) collected at a constant discharge of 142 m³/s.
- Table 2 Analysis on integrated sediment concentrations (g/L) collected at a constant discharge of 425 m³/s.
- Table 3 Irradiametric values calculated for a discharge of 142 m³/s. Coefficient values are listed for attenuation coefficients for scalar (K_o) , downward irradiance (K_d) , and net vertical attenuation (K_E) ; and also included are: coefficients for absorption, a, calculated at Z_m ; asymptotic backscattering (b_b) , normal backscattering (b_b) , and total scattering (b). The observed depths (m) for the compensation point (Z_{cp}) , were derived from either actual observed $in \ situ$ surface intensities (1), or calculations based on a constant incidental surface intensity of 2000 μ E.
- Table 4 Irradiametric values calculated for a discharge of 425 m³/s. Coefficient values are listed for attenuation coefficients for scalar (K_o) , downward irradiance (K_d) , and net vertical attenuation (K_E) ; and also included are: coefficients for absorption, a, calculated at Z_m ; asymptotic backscattering (b_b) , normal backscattering (b_b) , and total scattering (b). The observed depths (m) for the compensation point (Z_{cp}) , were derived from either actual observed $in \ situ$ surface intensities (I), or calculations based on a constant incidental surface intensity of 2000 μ E.
- Table 5 Discharge 142 m³/s, geomorphological characteristics for channel topography are calculated from STARS Simulation Model data, and correlated to irradiametric reach designations. The channel width, thalweg depth and slope data are derived from bed material maps supplied by USGS.
- Table 6 Discharge 425 m³/s, geomorphological characteristics for channel topography are calculated from STARS Simulation Model data, and correlated to irradiametric reach designations. The channel width, thalweg depth and slope data are derived from bed material maps supplied by USGS.

LIST OF FIGURES

Figure 1	Sampling site location on the Colorado River, Glen and Grand Canyons			
Figure 2	Light sensor deployment frame for measuring scalar and cosine corrected irradiance			
Figure 3	Concentration of suspended sediment (g/L) measured at a discharge of 142 m ³ /s			
Figure 4	Concentration of suspended sediment (g/L) measured at a discharge of 425 m³/s			
Figure 5	Regression demonstrating linear relationship between vertical attenuation ($K_{\rm o}$) and mean sediment concentration ($Q_{\rm s}$)			
Figure 6	Vertical attenuation coefficients for scalar irradiance (K_o) measured at discharges of 142 and 425 m ³ /s	18		
Figure 7	Absorptance coefficients (a) for the Colorado River, Glen and Grand Canyons measured at a discharge of 142 m³/s			
Figure 8	Asymptotic backscattering coefficient (b'_b) to nephelometric turbidity units (T_n)	23		
Figure 9	Longitudinal distribution of light attenuation in the Colorado River measured at a discharge of 142 m³/s			
Figure 10	Longitudinal distribution of light attenuation in the Colorado River measured at a discharge of 425 m³/s			
Figure 11	Vertical attenuation coefficients for cosine corrected irradiance (K_d) measured at discharges of 142 and 425 m ³ /s	54		

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Photosynthetically Available Radiation (PAR)

in the Colorado River: Glen and Grand Canyon

ABSTRACT

Photosynthetically available radiation (PAR) is a limiting factor on primary productivity and growth of aquatic algae in the Colorado River of Glen and Grand Canyons. The primary component contributing to light attenuation is suspended sediment which reduces the spatial extent of PAR in depth and distance downstream under varying discharges at Glen Canyon Dam. A longitudinally stratified sampling approach was used to characterize the optical properties influencing light attenuation. Our analysis indicates light attenuation increases from Glen Canyon to Diamond Creek and is significantly related to increasing sediment concentrations that occur with increasing distance downstream and discharges measured at 142 and 425 m³/s. At these constant discharges sediment amplification results from increasing downstream distances, fluvial hydrology and local geomorphology. Under high water clarity conditions with no sediment contribution from tributaries the compensation point for Cladophora glomerata ranges in depth from 19.4 to 3.6 m.

Acronyms and Symbols

a	= Absorptance coefficient	1	= Incidental light intensity (m ² /s)
b	= Total scattering coefficient	l _{ss}	= Sub-surface intensity (m ² /s)
b _b	= Normal backscattering coefficient	<i>T</i> ,	= Nephelometric turbidity unit (NTU)
b' _b	= Asymptotic backscattering coefficient	PAR	= Photosynthetically available radiation
E	= Net downward irradiance (E_d - E_u)	R_a	= Asymptotic reflectance
E _d	= Downward irradiance	a,	= Suspended sediment concentration (g/L)
E _u	= Upward reflectance	Z _{op}	= Compensation point depth
E.	≐ Scalar irradiance	Z.	= Euphotic zone depth (1% of PAR)
κ	= Vertical attenuation coefficient	Z _m	= Euphotic zone depth (10% of PAR)
Kavg	= Average vertical attenuation coefficient	Z _{sD}	= Secchi disc depth
K_d	= Attenuation coefficient for cosine irradiance	μΕ	= Microeinsteins (m 2 /s), (6.02 · 10 17 quanta)
K.	= Attenuation coefficient for scalar irradiance	K _E	= Attenuation coefficient for net irradiance

INTRODUCTION

Photosynthetically available radiation, *PAR* (400-700 nm wavelength), is critical to underwater photosynthesis and its subsequent productivity. Numerous factors which influence primary production and the proportional growth of algae in aquatic ecosystems include nutrient loads (Mantai 1978), temperature (Hodgson 1981), channel geomorphology (Tett *et al.* 1978), suspended sediment loads (Jewson and Taylor 1978), and seasonal light variation (Adams and Stone 1973; Graham *et al.* 1982). In this study we characterized spatially the light attenuation in the Colorado River, Glen and Grand Canyons, Arizona. The primary goal was to determine if under periods of minimal tributary discharge subsurface light availability was influenced by normal operations at Glen Canyon Dam.

Background Information

Prior to completion of Glen Canyon Dam in 1963, suspended sediment loads in the Colorado River in Glen and Grand canyons, exceeded 10,000 ppm (Dolan *et al.* 1974; Schmidt and Graf 1990). The operation of Glen Canyon Dam has since regulated flows and decreased sediment transport in the Colorado River (Pemberton 1976; Howard and Dolan 1981). The impoundment of Lake Powell and the resulting high clarity discharge abruptly shifted a previously allochthonous system of transported inorganic and organic material to an autotrophically based ecosystem. At present, tributary discharge of sediment into the Colorado River periodically shifts this relatively new riverine ecosystem into a quasi pre-dam environment.

Frequency, duration, and magnitude of tributary floods in the Grand Canyon are unpredictable, and their downstream extent and residency are quite variable (Webb 1987). These events can result in suspension of fine sands, colloidal silts, and clays which thereby partially or totally eliminate light penetration through the water surface. Suspended sediment loads are contributed to the river in three ways: 1) seasonal sediment input from perennial stream flow from the primary tributaries of

the Paria River, Little Colorado River, and Kanab Creek (Graf et al. 1990, Herford 1984); 2) infrequent flood and debris flow events from ephemeral drainage basins, (Webb 1987); and 3) degradation of alluvial deposits, (Schmidt and Graf 1990).

Problem Statement

Light availability becomes a critical aspect since the basal portion of the present aquatic productivity is derived primarily from photosynthesis (Usher *et al.* 1987; Hardwick *et al.* 1992). The degree of water clarity in the Colorado River and the resulting light penetration is inversely related to the presence of suspended sediment, and is functionally governed by sediment contribution from major tributaries, episodic events from ephemeral drainages and washes, and the degradation of alluvial deposits. The unpredictable and periodic nature of flood events is the primary factor limiting light availability in the Colorado River.

This research has primarily focused on secondary light attenuation, which results from degraded alluvium during periods of optimum water clarity conditions under normal operational discharges at Glen Canyon Dam. Irradiametric measurements were collected during periods of limited sediment contribution from tributaries, however, the proportion of light available for photosynthesis progressively decreased with increasing distances from Glen Canyon Dam. Scott (1978) conducted irradiametric studies in marine estuaries experiencing periodic sediment contribution and approached this problem by assuming that differences in measured attenuation coefficients, K, from the mean or optimum attenuation coefficient, K_{avg} , were proportionally related to increasing levels of suspended sediment. The observed changes from a base attenuation value K_{avg} was attributed to periodic input of sediment from adjacent river systems. Conceptually, the same assumption holds true for the Colorado River where increased light attenuation is the result of changes in sediment concentration. However, in our study changes in sediment concentration are related to the suspension and transport of degraded alluvium and not tributary flow.

Transmission of light and its vertical dispersal through water is directly influenced by the optical properties controlling light attenuation, absorptance and scattering (Di Toro 1978; Kirk 1980a; Kirk 1980b; and Kirk 1983). Concentration of suspended sediment, size distribution, shape, and refractive indices affect the orientation and distance that light travels through water, and ultimately its attenuation (Spinrad *et al.* 1978). This scattering aspect further intensifies the degree of light attenuation with depth (Kirk 1980b). Decreasing intensities with increasing depth are an outcome of both the absorptive (Kirk 1980b) and scattering characteristics of water (Kirk 1977; Kirk 1980a). Vertical attenuation coefficients increase in turbid waters as a result of the increased absorption and scattering properties of suspended sediment.

Information from this study will assist researchers investigating aquatic productivity in the Colorado River by bridging the span between light availability, assimilation and energy transformation through the series of trophic levels. In coordination with other research in progress this paper identifies additional factors related to primary production in the Colorado River.

Objectives

- Determine if vertical light attenuation is significantly correlated to suspended sediment, discharge volume, and distance from Glen Canyon Dam.
- Determine if photosynthetically available radiation, PAR, at variable discharges of 142 and 425 m³/s varies spatially throughout the longitudinal extent of the Colorado River, Glen and Grand Canyons.
- Determine the degree of scattering and absorptance to light attenuation for the
 Colorado River at variable discharges.

- Determine the depth zonation of available light intensities (μΕ) for <u>Cladophora</u>
 glomerata at varying discharges in the Colorado River downstream from Glen
 Canyon Dam.
- 5. Determine if vertical attenuation coefficients, K, derived for downward and scalar irradiance are either equivalent or can be made equivalent.
- 6. Develop a method to correlate secchi depth measurements, $Z_{\rm SD}$, to vertical light attenuation.

MATERIALS AND METHODS

Study Area

The study area is located on the Colorado Plateau of Northern Arizona, where the Colorado River becomes incised within the meanders of the Glen and Grand canyons, flowing through geological strata consisting primarily of sedimentary (sandstones, shales, and limestones) and metamorphic (schist and granites) material. The primary source of water for the Colorado River, in Grand Canyon originates now as a hypolimnetic discharge from Lake Powell Reservoir, at Glen Canyon Dam (0.0 km). The river descends 700 m in elevation from the dam to Lake Mead a total distance of 472 km. Sample site locations are relative to downstream distances (km) from Glen Canyon Dam, Fig 1.

Equipment and Data Collection

Three river trips were conducted in conjunction with two prescribed discharge tests (GCES-II 1990), May 20-30, 1991 at 425 m³/s (15,000 ft³/s); June 28-30 at 142 m³/s (5,000 ft³/s); and July 12-14 at 142 m³/s (5,000 ft³/s). The scheduled dates were selected for three reasons: 1) low probability of sediment contribution from tributaries; 2) constant discharge; and 3) ability to

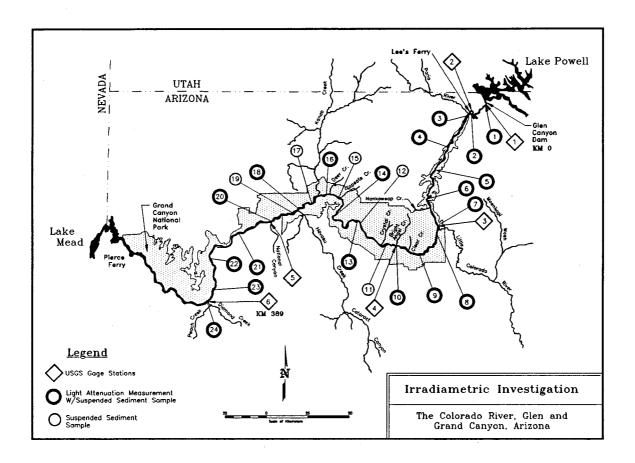
differentiate volumetric differences in discharge for 142 and 425 m³/s. Light attenuation measured at a specific discharge were not collected at all sites simultaneously, therefore, it can be assumed that these irradiametric measurements are sure to have been influenced to some degree by the antecedent conditions prior to or during each test flow.

Irradiametric measurements were collected using a series of underwater sensors (LiCor, Inc., Lincoln, Nebraska). Sensor types consisted of a spherical sensor (LI-193SA) collecting omnidirectional or scalar irradiance. Two cosine corrected quanta sensors (LI-192SA) were deployed at a 180° vertical orientation for measuring downward irradiance and upward reflected irradiance. A terrestrial (LI-190SA) quanta sensor was used for measuring incidental surface measurements of solar radiation. Each type of sensor measured the spectral region between 400-700 nm with equal sensitivity. A compatible data-logger (LI-1000) with multiple channel capabilities, collected and stored irradiant measurements. Irradiant units of measure are expressed in microeinsteins m^2/s , μE . One μE is equivalent to 6.02×10^{17} quanta/photon.

Adjustment for the immersion effect was accomplished using the appropriate multiplier specific to each photosensors air-water calibration setting (Roemer and Hoagland, 1979; Kirk, 1983). To avoid refractive problems, surface based measurements were collected at 0.025 and 0.07 m depths respectively for cosine corrected and scalar quanta sensors. Irradiance was measured within a 45° solar declination to account for reflectant loss by water surface. Research indicates that vertical attenuation coefficients are not overly influenced by daily shifts within this solar angle of incidence (Kirk 1977). For this reason measurements were restricted between 0900 and 1530 hr in order to minimize reflectant loss to ≤ 2.8% of the available solar radiation (Kirk 1983). To avoid hysteresis a 10 s sampling time allowed for sensor equilibration between depth adjustment. Selection criteria was established to bracket underwater irradiance simultaneously with measured incidental solar radiation. Profile measurements for underwater depth irradiance were clustered to

within a 15% range of the solar incidence to avoid problems associated with irradiant flux due to changes in atmospheric conditions.

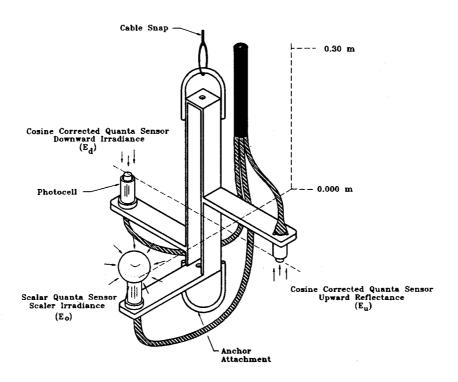
Figure 1 - The sampling site locations are numerically indicated for each irradiametric site conducted in the Colorado River, Glen and Grand Canyons.



Sensors were mounted independently on a frame consisting of multiple support arms constructed in such a fashion that all depth intervals were identical to each sensor type (Fig. 1.2). All light measurements were collected simultaneously and related to a specific depth. Sensor types were lowered operating a pulley system connected to a suspension boom over the sun-ward side of the boat. Measurements were collected at 0.5 m depth intervals in direct sunlight when weather permitted.

Figure 2 - Deployment frame used for collecting depth integrated measurements for cosine corrected and scalar irradiance. Each sensor type is positioned to collect simultaneous measurements for a specific depth. The cosine corrected sensors are vertically oriented to measure downward irradiance, $E_{\rm d}$, and upward reflectance, $E_{\rm u}$. Scalar irradiance, $E_{\rm o}$, is an omni-directional measurement (360°).

LIGHT SENSOR DEPLOYMENT ARM



Depth integrated samples for suspended sediment were collected at each site using U.S. Geological Survey's sediment sampling techniques. Sampling equipment consisted of a D-77 sampler mounted on a bridge boom. It was assumed that at a constant discharge an integrated depth sample of suspended sediment concentration and particle size distribution was representative of the cross sectional stream flow (Pemberton 1987; Einstein 1950). Three integrated samples were collected in midchannel along the thalweg for each irradiametric site. All samples were stored, transported and analyzed for particulates, ash free dry weight (AFDW) and sediment. Weights were obtained using glass filters (Whatman 934-AH 1.5 μ m pore size). Samples were desiccated for 1 hr. at 60°C, weighed (\pm 0.0001 g), and ashed for 3 hr. at 550°C, and reweighed (Guy

1969). The three samples collected per site were converted to a mean concentration (g/L) for particulates, particulate organic matter (POM), and sediment. These concentrations were analyzed against the calculated light attenuation coefficient for each sample site.

A Hach, Inc., turbidimeter (model 16800), was used to measure turbidity from each integrated sediment sample. Samples were measured in nephelometric turbidity units, T_n , defined as a 90° scattering of light by suspended particulates (Kirk 1980a; Kirk 1980b). Water quality data were collected using a Hydrolab Inc., DS2H, including temperature (°C), pH, conductivity (μ mhos/cm²), and dissolved oxygen (mg/L). A secchi disc was used to collect depth measurements in conjunction with irradiametric measurements at the same sample site.

Data Analysis

Photosynthetically available radiation, *PAR*, exponentially declines with depth (Kirk 1983; Williams *et al.* 1980). Light attenuation in water results from the combined effect of the physical components which scatter and ultimately absorb light, these include; water, soluble dyes (i.e. water-soluble humic substances), and suspended organic and inorganic material (Kirk 1977; Kirk 1983). Kirk (1977; 1980a; 1983) identified that the presence of inorganic and organic particulates are at times major factors in light attenuation for some natural bodies of water. The vertical attenuation coefficient, *K*, was used to characterize the Colorado River's light impeding properties as a function of discharge downstream from Glen Canyon Dam. This parameter allowed us to spatially characterize the availability of *PAR* as optical characteristics of the river system changed longitudinally with increasing distance from Glen Canyon Dam.

The exponential attenuation of *PAR* is linearized using eqn. 1, which is a natural log transformation of irradiance at a specific depth. The vertical attenuation coefficient is calculated

$$\ln Q = -Kz + \ln Q_o \tag{1}$$

using an equation developed by Kirk (1977; and 1983) which is a derivative of eqn. 2,

$$Q = Q_0 e^{-Kz}$$
 (2)

Refer to Appendix B, for additional information concerning methods used in calculating coefficients for vertical attenuation, absorptance, and scattering.

Optimum light conditions in the Colorado River for photosynthesis are best described as a relationship of depth where the compensation point of <u>C. glomerata</u> exceeds the channels mean thalweg depth. Under sub-optimum light conditions, channel depths exceed available light. Photosynthetic intensities (μ E) specific to <u>C. glomerata</u> were derived from previous research findings (Mantai 1974; Graham *et al.* 1982). As identified by Graham *et al.* (1982), the measured compensation point for <u>C. glomerata</u> extended from 25 to 35 μ E for water temperatures ranging from 5°C to 20°C. A mean intensity value of 30 μ E was selected as the compensation point for calculating $Z_{\rm ep}$ at variable incidental surface intensities rather then determining $Z_{\rm ep}$ from 1% *PAR*. Differences in saturation levels identified for <u>Cladophora spp.</u> vary between 920 μ E (Mantai, 1974), and 345 to 1125 μ E (Lester *et al.* 1974). Mantai (1974) reported that the point of photosynthetic saturation did not occur at low light intensities. We selected the depth of the saturation point based on the findings of Mantai (1974). The vertical zone for maximum net photosynthesis was calculated using a range in intensities from 300 to 600 μ E for each site (Graham *et al.* 1982).

The compensation depth was calculated using K the attenuating slope of subsurface light and the mean incidental light intensity measured during the depth profile by solving for the line-intercept. The equation (eqn. 3) below is used for determining the depth of the compensation point;

$$Z = -1/K \cdot \ln I + 1/K \cdot \ln I_{ss} \tag{3}$$

The derivative, eqn. 4, uses the constant 3.4 which is natural log transformation of the irradiant intensity identified as the compensation point, 30 μ E for <u>C. glomerata</u> (Graham 1982).

$$Z_{\text{compensation point}} = 3.4 - \ln I / K$$
 (4)

The constant (3.4) is specific to <u>C. glomerata</u> or other aquatic algae sharing similar compensation points and should be used as a relative base of measure if mean values of incidental light intensity are collected in conjunction with attenuation data (refer to Appendix B).

The sampling approach allowed us to characterize the optical properties of the Colorado River by stratifying the river longitudinally into irradiametric reaches. Sampling site locations were designated at 25 kilometer intervals, including specific sites located at major tributaries and the USGS gage stations. A total of 24 sites were sampled for suspended sediment concentration, and of these, 19 sites were correlated with irradiametric measurements. Locations of sampling sites on the Colorado River are indicated in Fig 1. This is an area that extends from Glen Canyon Dam to Diamond Creek, a distance of 387 km downstream.

Mean thalweg depths, velocities, channel slope and channel width for each designated reach were derived from data compiled from USGS bed material maps as part of the Sediment Transport and River Simulation (STARS) model (Randel and Pemberton 1987). The mean thalweg depth, channel slope and channel width were adjusted to reflect changes in vertical stage for discharges of 142 and 425 m³/s. For purposes of comparing geomorphology sampling sites were converted into 19 irradiametric reaches. The designated reach lengths were either half the distance between adjacent sites both upstream and downstream, or demarcated at the confluence point of the primary

tributaries (Paria River, Little Colorado River, and Kanab Creek). The calculated attenuation coefficient for each sample site is representative of this longitudinal distance. These hydraulic and geomorphological variables represent the channel characteristics for each of the designated irradiametric reaches. We compared this to 11 geomorphic reaches previously delineated by Graf et al. (1989), which represented general differences in the hydraulic characteristics and sedimentary features of the Colorado River. Our sample sites were overlaid with these distinct areas to determine if a relationship existed between geomorphology and vertical light attenuation. The channel geometry data were derived from the same source, however, the reach designations and flow discharges (682 m³/s) evaluated were different. For additional information regarding the hydrological characteristics of Grand Canyon reaches refer to Graf et al. (1989).

RESULTS

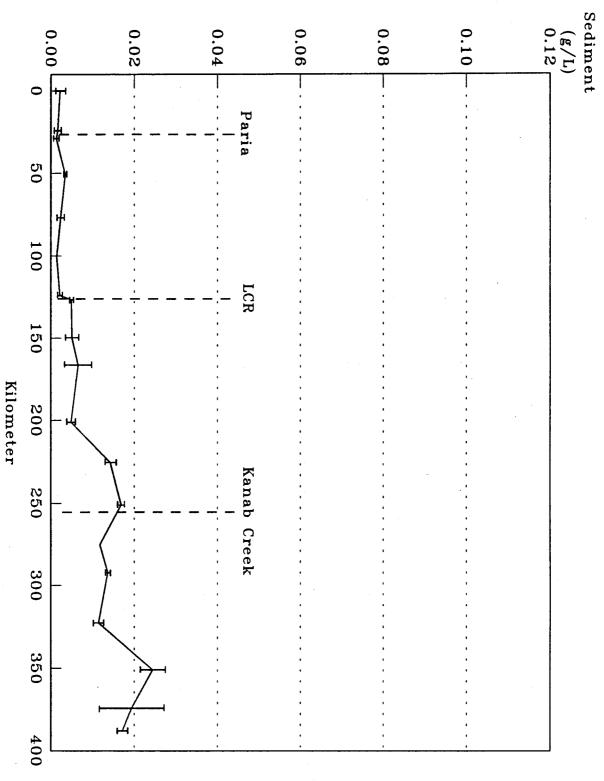
Analysis of Vertical Light Attenuation

The irradiametric data for the two evaluated discharges, 142 and 425 m³/s, were analyzed separately. It was identified that vertical attenuation coefficients for scalar irradiance are positively correlated to sediment concentration, distance and discharge from Glen Canyon Dam. Each of these factors demonstrated a positive correlation to light attenuation and are listed in order of greatest significance.

Sediment Concentration effects on Attenuation of PAR

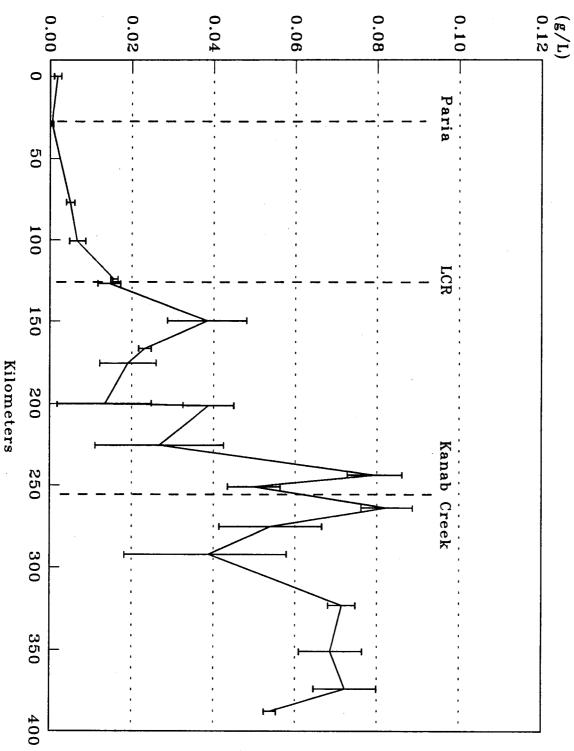
Concentration levels of suspended sediment varied at sites replicated at different discharges measured at 142 and 425 m³/s, (Fig. 3 and 4). The results from our analysis identified that light attenuation coefficients (K_o) were significantly correlated to particulates (i.e. this includes sediment and organic matter) at only the higher discharge 425 m³/s ($F_{1,18} = 42.432$, $R_{adj}^2 = 0.686$, p < 0.0001); however, no significance was found to exist at the lower discharge 142 m³/s. It was

by the error bar. Dam. The standart error in sediment concentration for each site is represented The mean concentration value is plotted against distance (km) from Glen Canyon Figure 3 - Concentration of suspended sediment (g/L) at discharge of 142 m³/s.



Dam. by the error bar. The mean concentration value is plotted against distance (km) from Glen Canyon The standard error in sediment concentration for each site is represented Concentration of suspended sediment (g/L) at discharge of 425 m³/s.





found that the fraction consisting of particulate organic matter (POM) collected as part of the integrated sample obfuscated our results. POM concentration was found not to be significantly correlated to light attenuation for either of the two discharges. The removal of organic particulates disclosed that light attenuation was significantly effected by sediment concentration at both discharges. The ANOVA's conducted on attenuation coefficients for scalar irradiance indicated a positive correlation between light attenuation and sediment concentration. The results for each discharge measured were significant ($F_{1,18} = 42.432$, $R_{adj}^2 = 0.686$, p < 0.0001) at 142 m³/s, and ($F_{1,17} = 75.192$, $R_{adj}^2 = 0.805$, p < 0.0001) at 425 m³/s. Elevated levels of suspended sediment were related to volumetric increases in discharge from Glen Canyon Dam as a result of accumulation in sediment load. The empirical data collected for sediment concentrations verifies the light attenuation data found in Table 1 and 2 (Appendix A).

The vertical attenuation coefficient, K_{\circ} , for scalar irradiance can be derived using data collected on mean sediment concentrations. The two developed regression equations for light attenuation are based on the compilation of mean sediment concentration data collected for all sites. The regression equations for calculating attenuation coefficients from variable sediment concentrations are identified: 1), for 142 m³/s discharge (eqn. 5);

$$K_o = 17.896 \cdot Q_s + 0.291 \quad (R^2_{adj} = 0.686)$$
 (5)

and 2), for 425 m³/s discharge (eqn. 6).

$$K_0 = 11.945 \cdot Q_s + 0.317 \quad (R_{adi}^2 = 0.805)$$
 (6)

The above results indicate a linear relationship between concentration and light attenuation for both regressions. The linear regressions were tested (students-t) to determine if significant differences

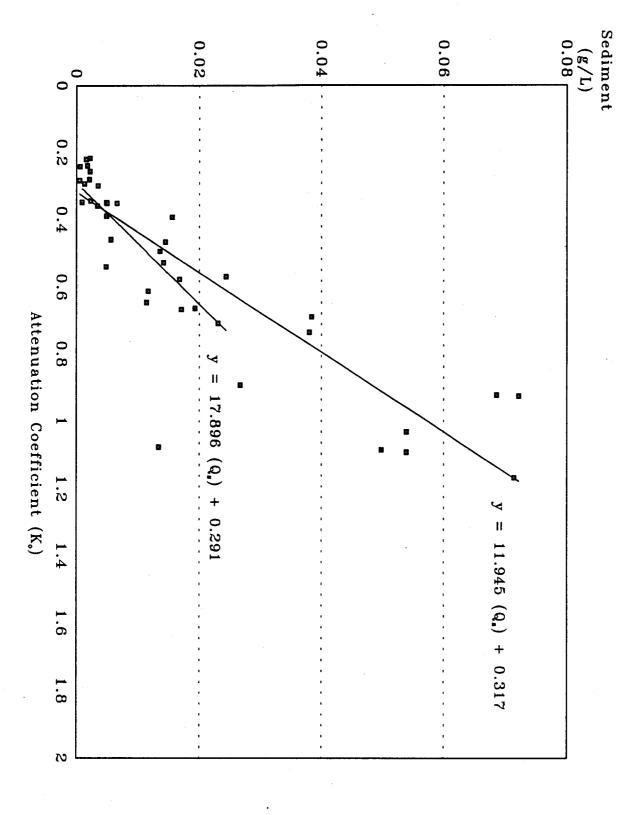
existed between the two calculated slopes. The findings indicate that both linear regressions are equivalent ($t_{0.05}$ (2), 35 = -0.057). Refer to Fig 5, for a graphic representation of the combined data set (n = 38) for both discharges. This graph demonstrates the linear relationship between vertical attenuation coefficients, K_o , and mean concentration of suspended sediment, Q_s . It appears that at this juncture, with the limited range of samples available that attenuation of light is a linear function to sediment concentration. However, from our analysis, the assumption that a linear relationship exists between sediment concentration and light attenuation over a wide range of sediment loads from discharges or tributaries in the Colorado River cannot be verified.

Distance effects on the Attenuation of PAR

As identified, distinct changes in light attenuation resulted with increasing distances downstream from Glen Canyon Dam. In evaluating light attenuation to increasing distances a positive correlation was observed to be significant regardless of the actual discharge volume. The results are listed respectively for discharges measured at 142 m³/s ($F_{1,18} = 150.551$, p < 0.001, $R^2_{adj} = 0.887$), and 425 m³/s ($F_{1,17} = 59.103$, $R^2_{adj} = 0.763$, p < .001). Refer to Fig 6, demonstrating a positive correlation between light attenuation and distance at discharges measured at 142 and 425 m³/s.

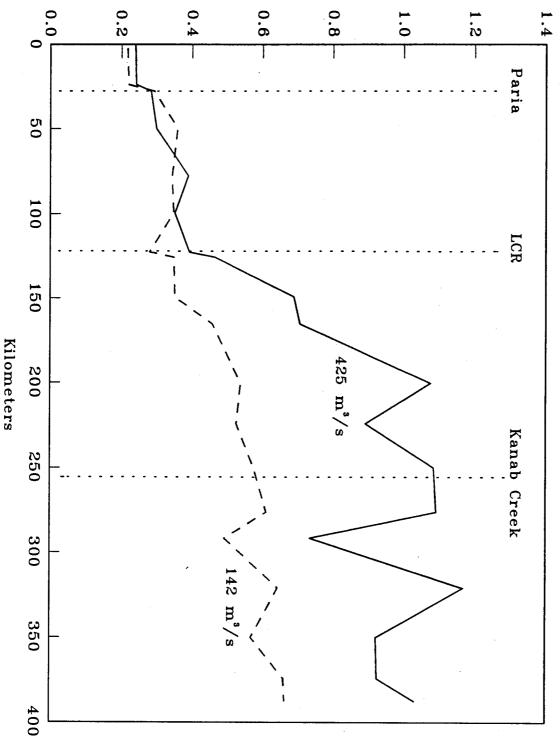
Vertical attenuation coefficients, K_o , were converted to percent light attenuation for the purposes of sample site comparison. The $K_{\rm avg}$, for Glen Canyon Dam represents the percent baseline (0%) for light attenuation in the Colorado River. The mean calculated attenuation coefficient, $K_{\rm avg}$, at the base of Glen Canyon Dam (0 km) during the summer period for scalar irradiance was 0.238 \pm 0.016 (s.d.) at discharges of 142 and 425 m³/s. This narrow range between attenuation coefficients is indicative of an optically stable source of water originating from Glen Canyon Dam within the discharge range evaluated during this experimental period. Differences in percent increase in light attenuation downstream from Glen Canyon Dam are based

 K_o , and mean concentration (g/L) of suspended sediment, Q_o . linear relationship between vertical attenuation coefficients for scalar irradiance, Figure 5 - The two regressions are equivalent and graphically demonstrate the



from Glen Canyon Dam (0 km) to Diamond Creek (387 km). at two steady state discharges of 142 m³/s and 425 m³/s on the Colorado River Vertical attenuation coefficients, (K_o), for scalar irradiance measured

Attenuation Coefficient



on the percent change from Glen Canyon's K_{avg} coefficient value. Percent change in light attenuation increased from Glen Canyon Dam to Diamond Creek a distance of 387 km. At Diamond Creek the maximum percent change in light attenuation varied from 180% (142 m³/s) to 335% (425 m²/s). This indicates that the differences in discharge volume influence the sediment transport capacity (Cluer 1992), and ultimately PAR by almost an order in magnitude. Also, percent change in vertical light attenuation with increasing distances downstream responded similar in pattern to the delineated hydraulic reaches of the Grand Canyon (Graf *et al.*, 1989).

It became apparent that the significance of distance (KM) as a variable effecting light attenuation was not independent of a combination of other interacting variables. In evaluating this phenomena, independent of the effect of discharge, we attributed the differences in light attenuation with increasing distances downstream to three factors; sediment carrying capacities, unique hydraulics and channel geomorphology of the Colorado River. At a constant discharge, the variation observed downstream in light attenuation were tested against the geomorphological changes in mean channel width, depth, and slope using a multiple stepwise procedure. In using a Pearson's correlation matrix on the collected and interpolated data developed from the Stars simulation model, we found no correlation between depth and slope with light attenuation. The only geomorphological variable that was found to be correlated to light attenuation was channel width, but only at a discharge of 425 m³/s. An ANOVA showed that a negative correlation for channel width to light attenuation was significant ($F_{1,16} = 7.166$, $R^2_{adj} = 0.278$, p = 0.017). The other geomorphic variables representing depth and slope appear not to significantly explain the response of increased light attenuation with increasing distances.

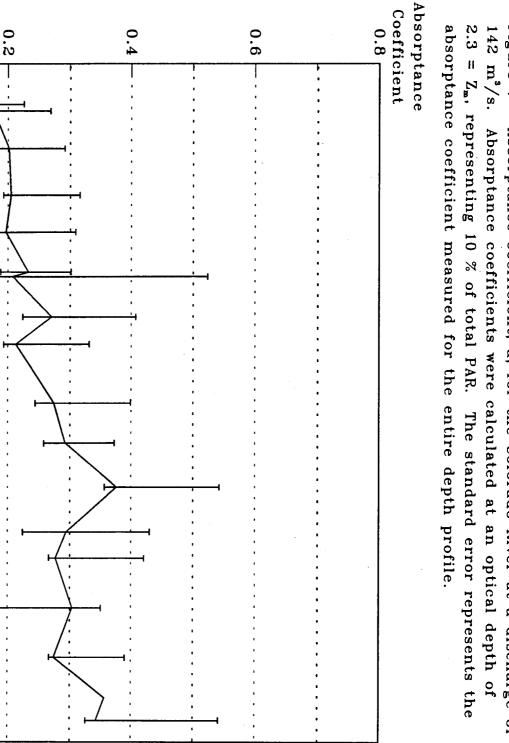
Light Absorptance and Scattering

Light attenuation is the direct result from the effects of both the absorptive and scattering characteristics of the constituents within the water column. Due to sampling difficulties

absorptance coefficients were derived only for a discharge of 142 m³/s. In analyzing the relationship of sediment to light absorptance it was determined that sediment was a significant factor ($F_{1,18} = 41.931$, $R_{adj}^2 = 0.683$, p < 0.001). In the Glen Canyon reach, light attenuation is minimal ($K_{avg} = 0.238$). Within this reach coefficient ratios of light absorptance, (a), to vertical attenuation, K, indicate that absorptance accounts for a higher proportion of the overall light attenuation, 71% ± 0.08 (s.d.). A distinct shift in light absorptance occurs below the Paria River at Site 3 (28 km). The shift in a/K ratios below the Paria tributary indicates that scattering becomes nearly equivalent to light absorptance in the overall aspect of light attenuation. The ratio of a/K remains fairly proportional at 55% \pm 0.05 (s.d.) for the remainder of the study area approximately 360 km, even though light attenuation for downward and scalar irradiance continues to increase downstream. It is speculated that the shift in a/K ratios are due to a change in optical characteristics below the Paria tributary as a result of the combined effect from light absorptance and scattering by sediment. Whereas, absorptance in the Glen Canyon Reach results from primarily water, organic particulates and dissolved dyes originating from Lake Powell reservoir. The light absorptance from these components becomes dominated by the overriding presence of sediment from the Paria downstream to Diamond Creek. Sediment accounts for the greatest aspect of light absorbance, and is graphically represented in Fig 7.

Total scattering (b) increases with distance downstream from Glen Canyon Dam. The varying values for scattering coefficients for the same site represent differences due to discharge volume indicating that scattering further increases light attenuation. The increased levels of suspended sediment disrupt the vertical orientation and directional path of light, further increasing the light scattering aspect of the water. Light scattering becomes very pronounced under higher discharges and in geomorphic reaches having higher turbulence. An ANOVA conducted showed significance $(F_{1,8} = 6.089, R_{adj}^2 = 0.361, p = 0.039)$ of sediment concentration to light scattering, with a positive trend. The absorptance and scattering properties of each site are presented in

 $142 \text{ m}^3/\text{s}.$ Absorptance coefficients were calculated at an optical depth of Absorptance coefficient, a, for the Colorado River at a discharge of The standard error represents the



Kilometers

Tables 3 and 4 (Appendix A), listing coefficient values for vertical attenuation, K, asymptotic reflectance, R_{\bullet} , asymptotic backscattering, $b'_{\rm b}$, normal backscattering, $b_{\rm b}$, and total scattering coefficient, b, for discharges of 142 and 425 m³/s. Nephelometric turbidity, $T_{\rm n}$, and the ratios between asymptotic backscattering and normal backscattering coefficients are also listed.

The correlation coefficient, R^2 , between asymptotic backscattering, b'_b , and turbidity, T_n , for all measurable sample sites at both discharges was 0.943. The mean ratio for T_n to b'_b was 32.5 \pm 4.9 (s.d.) in waters of varying turbidity. This reconfirms Kirks (1980a) research which produced a similar mean ratio value of 30.6 \pm 4.8 (s.d.) indicating that there is a linear relationship to asymptotic backscattering and turbidity (Fig 8). The asymptotic backscattering coefficient can be derived from nephelometric turbidity, T_n , applying eqn. 7.

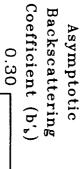
$$b'_{b} = 0.336 \cdot T_{0} - 0.00722 , (R^{2} = 0.943)$$
 (7)

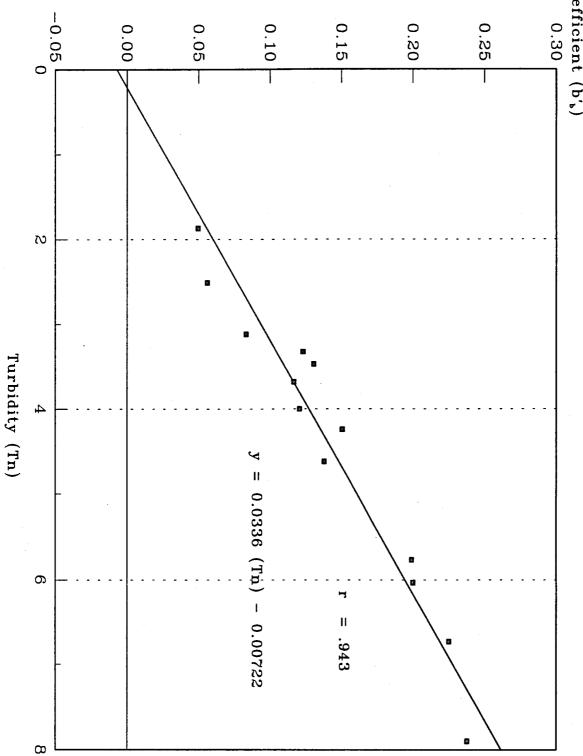
It was not possible to identify the asymptotic reflectance at all sites and discharge levels due to the affects of channel bed reflectance. In addition, R_a , was not encountered at certain sites because of profile depths in waters having a low K_d value.

Optical Light Conditions (µE) within the Colorado River's Euphotic Zone

The optical light conditions are defined by the underwater light regime relative to the compensation point for C. glomerata to the channels mean thalweg depth. Data generated from the STARS simulation model was reanalyzed to determine the effect of vertical change in stage on light availability. Our results indicate that for a volumetric change in discharge from 142 to 425 m³/s from Glen Canyon Dam to Diamond Creek results in an overall change in mean vertical stage of $1.68 \text{ m} \pm 0.55 \text{ (s.d.)}$, (n = 707). Table 5 and 6, lists mean thalweg depths, widths, slope, velocity, and compensation point and percent light availability for each site. The mean depth for

units (NTU). Figure 8 - Asymptotic backscattering coefficient, (b',), to nephelometric turbidity





142 m 3 /s is 5.26 \pm 4.13; and 425 m 3 /s is 6.95 \pm 4.32. The depths were adjusted to reflect differences in depth at these discharges and are representative of each irradiametric reach.

At a constant discharge of 142 m³/s the mean thalweg depth does not exceed the compensation point for C. glomerata for the entire extent of the Colorado River. Whereas, at a higher discharge (425 m³/s) sub-optimum light conditions predominate, restricting the compensation point (30 μ E) and the potential use of the channel bed to shallower depths for a large portion of Grand Canyon. The available area for algal growth will be reduced from the upper portion of the Granite Gorge (150 km) to Diamond Creek (387 km) at discharges of 425 m³/s or greater. At this discharge the mean thalweg depth exceeds the Z_{cp} , resulting in a 25 to 50 percent loss of the lower vertical zone of the channel. These reaches are subjected to photosynthetic exclusion or temporary concealment due to increased light attenuation. The increase in wetted perimeter resulting from a higher discharge (425 m³/s) will not compensate for the reduced photosynthetic zone in reaches of the middle and lower Grand Canyon due to increases in light attenuation. And yet, the optimum photosynthetically available light will exist from Glen Canyon to Little Colorado River under both discharges regimes at 142 and 425 m³/s. However, even though these optimum light conditions are available in the Marble Canyon and Lees Ferry reaches, the quantity in percent of total PAR will be appreciably reduced in irradiant intensity. Refer to Table 7 and 8 (Appendix A), for changes in percent change in total PAR.

Tables 7 and 8 (Appendix A), identify depths for observed and calculated Z_{cp} at flow discharges of 142 and 425 m³/s. The calculated depths for the compensation point, Z_{cp} , maximum net photosynthetic range (300-600 μ E), and the saturation point Z_{cp} are graphically represented in Fig 9 and 10. The zone of maximum net photosynthesis (300-600 μ E) is reduced from Glen Canyon to Diamond Creek at both discharges for the extent of the Colorado River.

Figure 9: Longitudinal distribution of light attenuation in the Colorado River at a 1420 m $^3/s$ discharge. Derived depths (m) for subsurface intensities are specific to \underline{C} , $g\underline{lomerata}$, from research identifying compensation point (Z $_{\pi}$), net photosynthesis (Graham et at. 1982), and saturation point (Mantai 1974). Mean channel depths have been adjusted to actual stage elevation (Randel and Pemberton 1989).

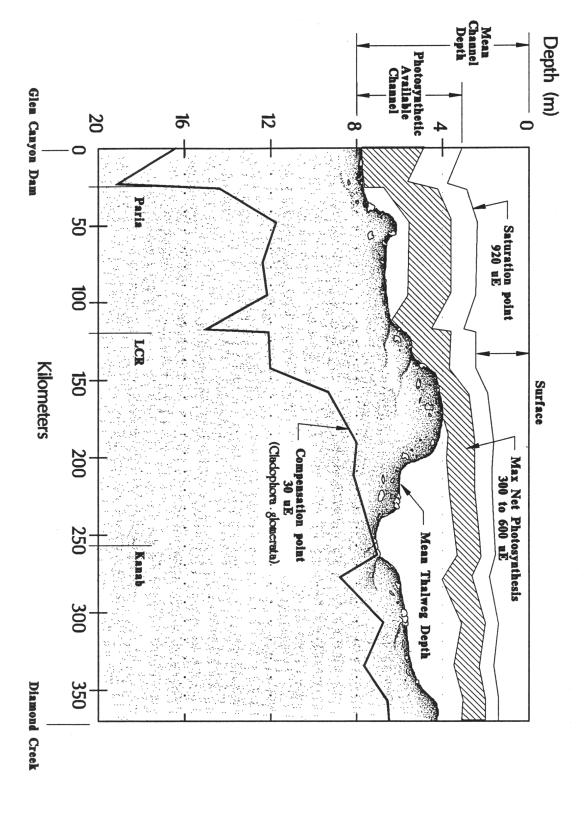
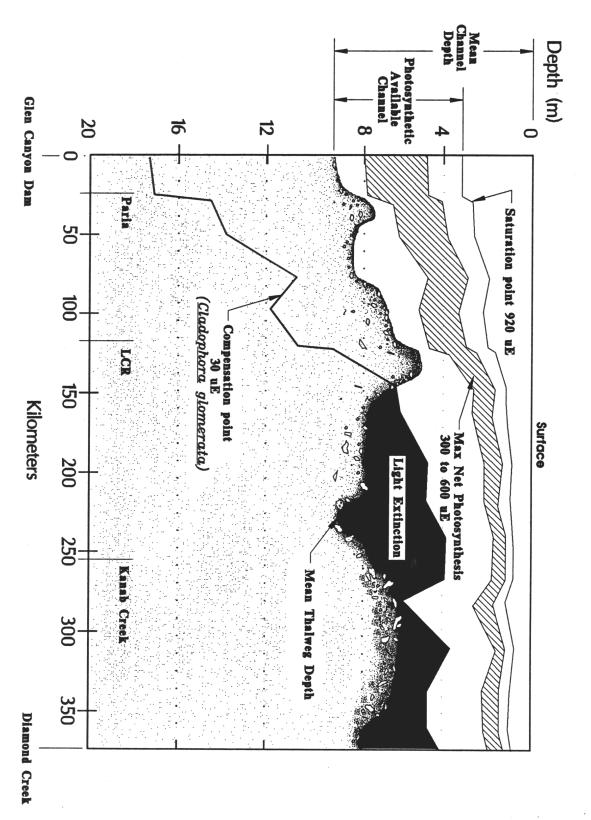


Figure 10: Longitudinal distribution of light attenuation in the Colorado River at a 425 m $^3/s$ discharge. Derived depths (m) for subsurface intensities are specific to \underline{C} , glomerata, from research identifying compensation point (Z $_{\pi}$), net photosynthesis (Graham et al. 1982), and saturation point (Mantai 1974). Mean channel depths have been adjusted to actual stage elevation (Randel and Pemberton 1989).



Light attenuation measured in the Glen Canyon area had similar attenuation coefficients under two constant discharges measured at 142 and 425 m³/s, refer to Tables 1 and 2 (Appendix A). However, on July 11, 1991, irradiametric measurements collected during fluctuating flow conditions for scalar irradiance showed an elevated attenuation coefficient of 0.302 ($R^2 = 0.990$) at a discharge level ranging from 736 m³/s to 793 m³/s (refer to the analysis section in Appendix C). Irradiametric measurements collected the following day at a constant discharge of 142 m³/s displayed reduced attenuation coefficient, K_o , of 0.220 , ($R^2 = 0.996$), which were similar to those previously measured attenuation coefficients for discharge levels at 142 and 425 m³/s. This indicates that discharges in excess of 425 m³/s have a disrupting effect on the stability of the underwater light regime as previously observed in the Glen Canyon area. Additional data has not been collected under conditions of fluctuating flows for other portions of the Colorado River downstream in Grand Canyon.

DISCUSSION

Analysis of Vertical Light Attenuation

Aquatic productivity derived from photosynthesis is strongly influenced by the underwater light regime throughout the Colorado River. Growth of <u>C. glomerata</u> is exceedingly important to the Colorado River since it represents a major biological substrate for epiphytyic diatoms (Hardwick *et al.* 1992). Difficulties arise in assessing the primary factors that control distribution and growth of aquatic plants due to the numerous variables encountered in a river system (Spence *et al.* 1970; Jewson and Taylor 1978). Certain studies have identified limiting factors that affect the photosynthetic capabilities of aquatic algae; these have focused on temperature optimums (Hodgson 1981; Graham *et al.* 1982), desiccation (Usher *et al.* 1987; Hodgson 1981), light intensities (Graham 1982; Marcus 1980; Jones and Ilmavirta 1978; Grande *et al.* 1989; Adams and Stone 1973; Spence and Chrystal 1970) and spectral composition of light (Pantastico and Suayan 1973; and Larkum *et al.* 1966). Usher *et al.* (1987), identified trends of decreasing

biomass in <u>C. glomerata</u> with increasing distances from Glen Canyon Dam, and speculated that seasonal growth may either be regulated by light or nutrient levels in waters of relatively constant temperature. Our data indicate that in the Colorado River light availability will become a primary limiting factor in aquatic productivity under a discharge regime of 425 m³/s or greater during periods of minimal concentration of sediment under tributary baseflow.

Sediment Concentration effects on the Attenuation of PAR

Our study has identified a relationship between light attenuation and sediment concentration under optimum water clarity conditions during periods of minimal tributary contribution. In other studies, methods used in correlating sediment concentrations to vertical light attenuation have not been entirely successful. This is partly attributed to the physical variability in particle size distribution, shape and refractive indices of the sediment (Spinrad 1978; Kirk 1980a). Also, it is speculated that our results showing a linear relationship are due to the removal of the particulate organic matter, since values for sediment concentrations generally do not account for this organic fraction.

Additional information exists on the effect of suspended sediment on light availability and its behavior (Kirk 1983; Jones and Wills 1956; Di Toro 1978; Duchrow Everhart 1971; and Scott 1978). The light attenuation regression (eqn. 6) provides researchers with an accurate estimator of light attenuation under a narrow range of sediment loads. Since the Colorado River can be extremely variable in sediment loads during periods of tributary influence our findings may not be applicable to correlating vertical light attenuation coefficients under excessive sediment concentrations (≥ 0.0722 g/L). As it stands, the linear relationships shown in eqn. 5, and eqn. 6, are not an accurate model for predicting vertical light attenuation derived outside of the sediment concentration range of this investigation. This is primarily due to the extreme variability in concentration, particle size and refractive differences of sediment transported from dissimilar tributaries (Kirk 1983; Randel and Pemberton 1987; Herford 1984). Therefore, additional analysis

for irradiametric and sediment concentrations are required at variable discharges and periods of tributary input to determine whether or not a predictive model can be developed using sediment concentration data. This information is sure to be beneficial if sediment concentrations are to prove to be a reliable predictor of light availability in the Colorado River.

The significant interference from particulate organic matter must be taken into account before attempting to correlate light attenuation to low sediment concentrations. The overriding interference is not significant at higher sediment concentrations. Additionally, it is interesting that our mean values for POM at 142 m³/s, 0.0081 ± 0.0012 g/L, n = 7; and at 425 m³/s, 0.0038 ± 0.0009 g/L, n = 5, collected in the Glen Canyon Reach were considerably higher than the mean values (0.0006 ± 0.00004 g/L) observed by Angradi (1992). The discrepancies in our findings are probably related to methodological differences (i.e. Miller-tube and AFDW at 2 h. intervals), sampling time and size. It is important to note that POM was observed to increase with increasing distances downstream, (refer to Table 1 and 2). We contend that the increase in POM is in part related to the export of POM from the Glen Canyon Reach in addition to photosynthetic activity occurring downstream.

Distance Effects on the Attenuation of PAR

As identified, increases in volumetric discharge from Glen Canyon Dam of 142 to 425 m³/s result in an overall increase in light attenuation by almost an order of two over a distance of 387 km.

Reaches where channel gradients decrease in slope, light attenuation became reduced but not in proportion to other reaches of similar hydraulic and geomorphic characteristics observed further upstream.

Decreases in mean channel depths and widths may explain the abrupt or gradual shifts observed in light attenuation at distinct sites or broad geographical areas. The reduction in channel depth can

resuspend sediment by increasing turbulence (Colby 1961); whereas, wide channel widths can influence flow characteristics by reducing velocities and the sediment carrying capacities. In such areas, reduced rates of erosion are indicative of areas with large channel widths (Dolan *et al.* 1974). Blinn *et al.* (1992) identified that wide reaches exhibited extremely high total standing crop percentiles in comparison to narrow reaches that were evaluated. We identified that a significant negative correlation exists with channel width to sediment concentration at higher discharges. This indicates that PAR is more available at greater depths in regions having larger channel widths which further supports Blinn's results.

The sediment transport capacity of a discharge dictates the concentration levels, and once sediment particles are actively resuspended, response time is not instantaneous to changes in hydraulic conditions. As documented by Cluer (1992) the sediment transport capacity at 425 m³/s was approximately 1,300 tons per diem, however, considerably lower then 50% of the experimental test discharges evaluated in the GCES Phase II study.

The suspension of fine sands, silts, and clays from periodic tributary discharges or pre-dam alluvium appears to be cumulative, and once inclined sediment remains in suspension (Laursen and Silverston 1976). This cascading effect maintains sediment in suspension and appears to be related to the fluvial dynamics of each study reach (Graf *et al.* 1989). The sediment transport capacity of the Colorado River and the progressive accumulation of sediment downstream appears to explain the response of increased light attenuation with increasing distances, since the independent geomorphological variables of channel depth, slope and to some extent width appear not to be significant. Also, the lack of an existing correlation to geomorphology may be a factor in attempting to characterize the channel geometry as a mean value representative of a 25 km distance. And yet, using interpolated data that was site specific or at least adjacent to the irradiametric sampling site showed no additional correlation.

Distinct shifts in light attenuation are related to point sources of sediment input from tributaries. The sediment storage differences in the river channel downstream from these confluence areas in addition to other catchment areas may explain these attenuation shifts. As discussed, tributaries will influence the availability of sediment, especially since both the Paria River and LCR contribute 72% of the Colorado River's annual sediment budget (Randel and Pemberton 1987). In addition, differences in water chemistry from the LCR tributary during base flow of 6.5 m³/s may account for further light attenuation below the LCR (Johnson and Sanderson 1968). The differences in percent change in light attenuation are related to discharge volume, hydraulics, and the availability of alluvial sediment (i.e. bank and channel storage).

As observed, the percent light attenuation at specific sites displayed distinct patterns of attenuation for both discharges. The mirror response at each site suggests that the hydraulic characteristics are unique and site specific. This response pattern in light attenuation is consistent regardless of the different discharges. A dramatic decrease in percent light attenuation occurred at the National Canyon site (20). This accounted for a net change of 50 and 150% decrease from the previous sampling site 15 km upstream for discharges at 142 and 425 m³/s. It is speculated that in this general area the hydraulic conditions are not conducive for maintaining suspended sediment in the water column.

Light Absorptance and Scattering

Light attenuation equals the sum of the optical properties of scattering and absorptance (Kirk 1983). Due to certain optical conditions loss of irradiance to reflectance may account for a significant portion of the available light for photosynthesis. The quantity of photosynthetically available radiation, *PAR*, is dependent on the presence and physical interaction with particle size, shape, concentration and refractive characteristics which influence the optical properties of the water (Kirk 1980b; Kirk 1983). It is speculated that the primary components of absorptance from

the hypolimnetic releases of Lake Powell's Reservoir are water, dissolved organic and inorganic compounds. In moderate to turbid waters scattering at all angles encompassing the spectrum of visible light (400-700 nm) is primarily due to particles (Kirk 1983). Normal backscattering, b_b , and total scattering coefficients, b, are the most useful types of information for characterizing the scattering properties of water (Kirk, 1980a). Variation in sediment particle size and concentration can significantly affect the optical properties of normal backscattering (Spinrad *et al.* 1978; Kirk 1980a; Kirk 1977). In the Colorado River the optical properties vary both spatially and temporally in response to normal operational discharges, distance downstream from Glen Canyon Dam, and the frequency of sediment contribution from tributaries.

Light attenuation is in part due to the hydraulic characteristics that increase and maintain suspended sediment with increasing distances downstream from Glen Canyon Dam. The progressive increase or abrupt shifts in light attenuation coincide with hydraulically distinct reaches, (refer to Graf *et al.* 1989). In addition, observed differences in light attenuation are attributed to the contrasting differences in sediment storage above and below the Paria and the LCR tributaries. Since 1963, degradational loss of alluvial material has occurred in the Glen Canyon area (Pemberton 1976). Whereas, sediment contribution from the Paria and LCR tributary's has continued to supplement the bed channel loss in sediment (Randel and Pemberton 1987). It is speculated that the sediment storage differences in all reaches below major tributaries account for the additional increases in observed light attenuation.

Our data on absorptance/attenuation (a/K) ratios indicate that the primary constituent responsible for light absorptance in the Glen Canyon area is not sediment. We contend that the shift in ratios of a/K as observed below the Paria tributary are due to sediment differences and availability of sediment both in channel storage and alluvial beaches below major tributaries. Otherwise, the a/K ratios should have remained constant if there was a proportional increase in all attenuating

components. However, a/K ratios that shift may account for a nonproportional increase of one of the many components responsible for light attenuation; i.e, soluble dyes, POM and sediment. In addition, sediment data support our contention that the primary component involved in light attenuation below the Paria tributary and progressively downstream is suspended sediment.

The degree of light scattering is dependent on size and concentration of suspended inorganic and organic particulates. Coefficients of normal backscattering have been used to determine suspended sediment concentrations based on the relationship that light attenuation equals the sum of absorptance and normal backscattering, b_b. Studies on the Yellow River in China, using remote sensing methods (Landsat MSS) have identified a correlation between empirical data collected on sediment concentrations with similar estimates using backscattering reflectance (Aranuvachapu and Walling, 1988). In certain situations a linear relationship occurs with suspended matter and the scattering coefficient (Jones and Will 1956). In remote sensing studies, a linear relationship with suspended sediment concentration and scattering is obtained by isolating the red and near infrared waveband. Absorption in the red waveband remains fairly constant, this is principally due to the fact that water is the primary medium of absorptance for this spectral region (Klemas et al. 1973; Aranuvachapu and Walling 1988). A linear relationship with suspended sediment will occur only if the absorption coefficient remains constant since reflectance varies inversely with absorptance while increasing with b_b at the 400-700 nm bandwidth (Kirk 1983). However, this method would be impractical for determining sediment concentration due to limitations in river size and background interference from shoreline margin.

Conditions of differing levels of turbidity can influence the spectral distribution of irradiance in addition to the availability of total *PAR* (Kirk 1979). Scattering increases in proportion to concentration with a strong indication that within the visible spectrum of light (400-700 nm) the shorter wavelength becomes easily scattered (Morel 1973). Under minimal tributary flow, non-algal

components either dissolved or colloidal compounds are of an allochtonous source and are generally minimal in concentration; however, their presence can reduce the quality and quantity of PAR for photosynthesis (Kirk 1980a; and Jewson and Taylor 1978). Partial or total elimination of different wavelengths of visible light can influence algal composition and vertical distribution. The adaptive capabilities of different algal species to utilize both the changes in the spectral region and light intensities will influence algal composition and their distribution (Larkum *et al.* 1966). Blinn *et al.* (1992) identifies that algal dominance shifts to favor <u>Oscillatoria</u> downstream from the Little Colorado River. It is speculated that the shift in composition is an adaptive response by this species to tolerate desiccation, light availability and the spectral shift away from the blue waveband.

The optical characteristics of water classified by Kirk (1980a and 1983) indicate that waters with a large particulate fraction will absorb light more strongly than the soluble fraction. Typically, under conditions of high clarity and low soluble dyes, water will have a higher affinity for absorbing the longer waveband (600-700 nm) of the orange-red region. The visible light spectrum shifts in natural waters containing high concentrations of phytoplankton and favors a greater penetration of the red region (Kirk 1979). At this point in time plankton densities in the Colorado River are of such low concentrations that they are considered by us to be an insignificant component of the overall constituents involved in light absorption (Haury, 1986). In conditions of high turbidity the shorter wavelength in the blue (400-500 nm) or blue-green (400-550 nm) region are attenuated in the upper surface (Kirk 1983). This spectral shift eliminates or reduces the utilization of this very important spectral region (400-500 nm). This has negative implications for growth and photosynthetic yield of <u>C. glomerata</u> which utilizes this spectral region (Kirk 1983).

Optical Conditions within the Colorado River's Euphotic Zone

The attenuation of underwater light directly influences the spatial quantity and density of algae

present within the euphotic zone, (Ganf 1974; Jewson and Taylor 1978), gross photosynthetic yields per unit area (Vollenweider 1960), and the temporal rate of growth (Jewson Taylor 1978). Literature supports an adaptation by <u>C. glomerata</u> to low light intensities (Mantai 1974; Lester *et al.* 1974; Adams and Stone 1973). Therefore, a spatial relationship for light attenuation exists as a function not only of depth, but distance and discharge from Glen Canyon Dam.

It can be surmised that periods of exclusion or concealment below the compensation point is another factor controlling vertical zonation of this algae. Resident time of algae below this lower limit restricts growth rate, photosynthetic yield, and areas for colonization (Jewson and Taylor 1978). The time of residence for phytoplankton in non-illuminated zones in turbid lakes has been shown to influence net yields of photosynthesis per unit area and growth (Jewson and Taylor 1978). Increased light attenuation will reduce or totally eliminate certain areas that might otherwise be available for colonization and growth (Spence and Chrystal 1970). This becomes further complicated once you consider the temporal (diel and seasonal) duration and variability of the underwater light regime in the Colorado River. For this reason, it is speculated that colonization and growth are restricted closer to the channel margin during suboptimum light conditions in areas located 150 km downstream or greater from Glen Canyon Dam. In the lower portion of the Grand Canyon this reduction in depth has negative implications for <u>C. glomerata</u> because fluctuating flows may potentially subject the only available area for attachment and growth to the effects of desiccation.

In a different light, a significant factor is exposure time, primarily because of the sessile nature of <u>C. glomerata</u>. Under fluctuating flow regimes, loss in surface area due to channel depth reduction is further compounded by exposure above saturation point potentially resulting in photoinhibition (Mantai 1974; Adams and Stone 1973; Boston and Hill 1991), and desiccation (Usher *et al.* 1987). Studies on exposure of this algae indicated a depreciable effect within the intertidal zone from

desiccation due to fluctuating flow patterns generated by Glen Canyon in the Colorado River (Usher et al. 1987). Similar studies have identified effects from desiccation, temperature and light intensities brought about by exposure (Hodgson, 1980; Adams and Stone 1973; and Pantastico and Suayan 1973).

In the Lees Ferry area growth and maintenance of <u>C. glomerata</u> requires a minimum 0.5 m subsurface depth regardless of flow regime (Pinney 1991). Besides the effect of desiccation, this observed depth response may be related to areas exposed at or above the saturation point. Exposure to the red light spectrum has been identified with aging and fragmentation of <u>C. glomerata</u> (Pantastico and Suayan 1975). Mantai (1974) reported that the photosynthetic saturation point occurred at light intensities of 920 µE. Information is not available on intensity levels and duration required to initiate a photoinhibitive response in <u>C. glomerata</u>, however, photoinhibition above the point of light saturation will reduce the photosynthetic rate in all aquatic algae. Research on photoinhibition of phytoplankton, <u>Asterionella</u>, indicates an exponential decline in rate of recovery to the duration of exposure (Kirk 1983; Boston and Hill 1991). Since increased exposure can offset recovery rates subsurface depths with high intensities could limit the upper vertical distribution of <u>C. glomerata</u> in waters having low light attenuation as found in the Glen Canyon area.

Optimum temperatures for photosynthesis of <u>C. glomerata</u> range slightly higher, 13 to 17°C, then the temperatures at Lees Ferry, 8 to 12°C (Pinney 1991). Graham (1982) reported that growth increased in colder temperatures with higher light intensities under controlled laboratory conditions. The inverse was true for warmer temperatures. The maximum net photosynthesis occurs at water temperatures ranging from 13° to 17°C, for light intensities ranging between 300 to 600 μ E (Graham *et al.* 1982). During our study the observed temperatures for the entire length of the Colorado River ranged from 8.7 to 17°C (142 m³/s), and 8.5 to 13.3°C (425 m³/s). The

and mass of water for each distinct discharge, and its downstream transit time. Due to the colder stenothermic releases from Glen Canyon Dam (Ward and Stanford 1979), the calculated zones for maximum net photosynthesis and the saturation point may slightly overestimate the actual photosynthetic range (Graham *et al.* 1982) for the upper portion of Glen and Grand canyons. And yet, in the middle and lower portions of the Grand Canyon water temperatures are at an optimum for <u>C. glomerata</u>, and may reflect the actual euphotic zone.

Glen Canyon Dam Operations

As identified, discharge volume directly influences the degree of light attenuation and is an independent variable that is directly controlled by operational management decisions. It was identified that variable discharges increases light attenuation in the Lees Ferry area. Certain research flows at varying ramp rates at Glen Canyon Dam were identified as causing greater degradation of alluvial deposits (GCES-Interim Flow Recommendations 1991). If our measurements are an indication of how light attenuation increases above 425 m³/s in the Glen Canyon reach that is known to be sediment depleted, light attenuation should increase dramatically downstream (Pemberton 1976). The "stair-step" decrease in standing crop of primary producers and macroinvertebrates observed in the Colorado River as associated with the primary tributaries indicates the limiting effect on primary and secondary productivity by sediment (Blinn *et al.* 1992; and Usher *et al.* 1990). Furthermore, it becomes apparent that under a fluctuating flow regime, rapid ascending or descending limbs of the hydrograph would display greater light attenuation than at a constant discharge of similar volume. It is speculated that additional light attenuation would be evident at similar discharges (142 and 425 m³/s) if ramping rates were factored into this equation (Cluer 1992).

Secondly, operational management of Lake Powell reservoir could further accentuate the observed

light attenuation downstream. The reservoir usually exhibits an extended period of stratification and convective mixing during the winter period, and generally does not extend for the entire reservoir depth (Stanford and Ward 1990; Johnson and Merritt 1979). However, under lower lake levels the withdrawal depth as related to surface elevation of Lake Powell may influence water quality and the attenuating characteristics of that discharge release. Seasonal variation in water clarity of Glen Canyon Dam releases would then result in greater light attenuation. We identified a distinct shift toward greater light attenuation during the winter period. On 16 December 1991, during a period of lake mixing (per comm. Verneiu), irradiametric measurements collected showed an elevated attenuation coefficient of 0.302 (r^2 , 0.955) for scalar irradiance (refer to Appendix D). Preliminary data indicates that lower water elevation in the reservoir combined with lake turnover may be responsible for elevated vertical attenuation coefficients. If this water quality condition is prevalent during winter periods the compensation depth would be further reduced, this is also in addition to the reduction in seasonal solar intensities and exposure time expected during this same period. This indicates that the use of a constant K_{avq} for seasonal comparisons is only applicable during conditions conducive for stratification of the reservoir prior to the turnover of the epilimnion.

CONCLUSIONS

Light attenuation is positively correlated to both increasing discharges and distances downstream from Glen Canyon Dam. Suspended sediment is the most significant factor regulating PAR in the Colorado River, Glen and Grand Canyons. The major factors influencing light attenuation are as follows: 1) sediment discharge from major tributaries; 2) discharge releases from Glen Canyon Dam; 3) sediment storage differences below major tributaries; and 4) channel geometry influencing fluvial hydrology. As speculated by Pinney (1991), seasonal changes in light intensity coupled with fluctuating flow levels are the most regulating factors for <u>C. glomerata</u> biomass in Glen Canyon. At present limited information is available on the seasonal shift and duration of exposure of *PAR* and frequency of sediment discharge from tributaries into the Colorado River.

The light absorptance by sediment appears to be a major factor influencing vertical light attenuation. Light scattering by the reflectant properties of sediment increases with downstream distances. Both of these mechanisms in light attenuation have negative implications for the quantity and spectral composition of light available for photosynthesis. Optimum light conditions for photosynthesis are determined by the optical properties of water and exist when the compensation point depth for C. glomerata exceeds the channels mean thalweg depth. It was found that at lower discharges (142 m³/s) the depth of the river channel does not exceed the lower limits of C. glomerata compensation point for the entire length of Glen and Grand Canyons. These optical conditions are optimal for photosynthesis. However, at higher discharges (425 m³/s) the depth of the river channel exceeds the lower limits of C. glomerata compensation point. This reduction in vertical depth generates suboptimum light conditions for photosynthesis from the Granite Gorge 150 km to Diamond Creek 387 km. The quantity and availability of light is influenced by the duration of optimum light conditions, frequency between periods of tributary discharges, and the spatial differences in light attenuation downstream from Glen Canyon Dam. The reduction in solar intensities and duration will decrease the total PAR available for aquatic productivity.

In addition, the use of other light measuring methods have applications in characterizing PAR since correlations have been established with these other methods (e.g. cosine corrected sensors and secchi disc). Vertical attenuation coefficients derived from irradiametric methods using either cosine corrected or scalar sensors are significantly different, but can be made equivalent. Also, secchi depth measurements collected in the Colorado River, Glen and Grand Canyons can be used to derive vertical attenuation coefficients, K_{\circ} , for scalar irradiance. Refer to Appendix B for further information regarding techniques for correlating other light measuring techniques.

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APPENDIX A

TABLES

PHOTOSYNTHETICALLY AVAILABLE RADIATION (PAR)

IN THE COLORADO RIVER: GLEN AND GRAND CANYON

			L) 0.008714 0.007693 0.001001	Mean Sediment (g/L) Sediment (s.d.) S.E.	Sed S.E.		0.007497 0.003825 0.000471	Mean Organic (g/L) Organict (a.d.) S.E.	S E				not available (*)	Data not ava
			0.0180		;	:	0.0141		:		0.0320	387.9	910702	5K24G
			0.0143				0.0056				0.0199	387.9	910702	5K 24B
0.001184	0.0021	0.0171	0.0191	0.002070	0.0036	0.0092	0.0078	0.002800	0.0050	0.0263	0.0269	387.9	910702	5K 24A
			0.0279				0.0105				0.0444	374.2	910702	5K23C
0.007733	0.0134	0.0193	0.0004	0.003030	0.0052	0.0100	0.0036	0.010377	0.0180	0.0293	0.0040	374.2	910702	5K 23A
			0.0184				0.0103				0.0288	350.0	910702	5K 22 C
0.002933	0.0051	0.0244	0.0240	0.001603	0.0028	0.0130	0.0109	0,004508	0.00/8	0.0375	0.0477	350.0	010702	5K 22B
			0.0108				0.0048				0.0150	322.7	910701	5K 21C
		!	0.0002				0.0080				0.0172	322.7	910701	5K 21B
0 001170	0 0000	0.0114	0.0141	0.001758	0.0030	0.0084	0.0122	0.002720	0.0047	0.0197	0.0263	322.7	910701	5K21A
			0.0124				0.0054				0.0178	202.2	910701	5K 20 E
0. 000506	0.0009	0.0136	0.0144	0.001501	0.0026	0.0091	0.0108	0.002001	0.0035	0.0227	0.0253	292.2	910701	5K.20A
			0.0139				0.0077				0.0216	275.3	910630	5K 18C
0,000956	0.0017	0.0117	0.0114	0.000715	0.0012	0.0062	0.0004	0.001860	0.0029	0.0180	0.01/8	275.3 275.3	910030	5K 18B
-			0.0182				0,0066				0.0248	251.1	910630	5K 18 C
6. 666 7 7 5	0.00	0.00	0.0150		9.00		0.0053				0.0202	251.1	910030	5K 10B
0 000 776			0.0132	00000		999	0.0084	0 001977	0 0000	0 0233	0.0250	251.1	910630	5K 16A
			0.0174				0.0121				0.0295	225.4	910636	5K 14B
0.001295	0.0022	0.0142	0.0121	0.000886	0.0015	0.0103	0.0103	0.002034	0.0035	0.0245	0.0225	225.4	910030	5K 14A
			0.0046				0.0038				0.0050	201.2	910029	5K 13C
0.000993	0.0017	0.0048	0.0070	0.000272	0.0005	0.0035	0.0038	0.001226	0.0021	0.0083	0.0108	201.2	910029	5K 13A
			0.0022				0.0026				0.0048	100.0	910715	5K 10C-2
0.0031/4	0.0055	0.0065	0.0030	0.000286	0,0000	0.0031	0.0038	0.003401	0.0000	0.0000	0.0181	186.0	910715	5K 10B-2
			0,0061				0.0079				0.0139	1 6 G	82001B	5K 10 C
			0.0085				0.0117				0.0181	160.0	910629	5K 10B
			0.0087	0.000995	0.0017	0,0093	0.0082	0.001397	0.0024	0.0148	0.0124	160.6	910629	5K 10A
			0.0042				0.0042				0.0084	140.7	910715	2 S
0.001552	0.0027	0.0050	0.0022	0.000747	0.0013	0.0026	0.0010	0.001917	0.0033	0.0076	0.0032	149.7	910715	5K 9A
			0.0048				0.0080				0.0139	126.7	910714	5K 8C
0.000430	0.0007	0.0049	0.0040	0.000470	0.0008	0.0083	0.0094	0.000377	0. 900 7	0.0132	0.0134	126.7	910714	5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5
			0.0028				0.0118				0.0140	124.0	910714	SK7C
		,	0.0020	9		!	0.0076	;	į		0.0103	124.0	910714	5K 78
0.000526	0.0009	0 0021	0.0008	0.001016	0.0018	0.0094	0.0086	0.001310	0.0023	0.0114	0.0094	124.0	910714	5K 7A
			0.0014				0. 008 1				0.0000	50.0	919713	2K 08
0.000000	0.0000	0.0014	0.0014	0.002373	0.0034	0.0044	0.0050	0.003234	0, 0056	0.0119	0.0004	100.0	910713	5K BA
			0.0002				0.0077				0.0107	77.0	910713	5K5C
0.000863	0.0015	0.0023	0.0036	0.000350	0,0006	0.0076	0.0083	0.001215	0.0021	0.0099	0.0119	77.0	910713	SK SA
			0.0026				0.0018				0.0044	50.8	910713	5K.4C
0.000328	0.0008	0.0034	0.0038	0.000357	0.0000	0.0023	0.0032	0.000615	0.0011	0.0007	0.00/0	00 C	010713	5X 48
			0.0016				0.0000				0.0076	29.0	910712	5K 3C
0.00000	0.000	0,0010	* 0022	6. 665 617	9. 86 40		• !		!		0.0042	29.0	910712	5K.38
			0.0020	00014	9	3	0.0052	0 00 1900	0 0034	0 008 1	0.0072	2 24	910712	5K2C
	:	!	0.0028				0.0042				0.0070	24.2	910712	5K 2B
0.000832	0.0012	0.0010	0.0045	0.001593	0.0023	0.0031	• 0130	0.000078	0.0012	0, 0063	0.0040	24.2	910712	5K2A
			•				•				0.0040	9 0	910712	5K 1B - 2
0.001298	0.0018	0.0022	0.0020	0.004417	0.0062	0,0088	0.0130	0.003236	0,0056	0.0126	0,0150		910712	5
			0.0024				0.007				0.0079	0 !	929628	9. ic
0.000338	0.0006	0.0022	0.0028	0,000148	0.0003	0, 0069	0.0070	0.000482	0. 0008	0. 009 1	0.0000	9 0	910628	5K 1A
	•.d.	9/1	9/2		• •	9/2	97		•	9/2	9/2	×	DA IE	SAMPLE NO.
ERROR	CONC.	CONC.	CONC	ERROR	CONC	CONC	CONC	ERROR	CONC	CONC	CONC.		!	
MEAN	SEDIMENT	MEAN	PERMIT	MEAN	ORGANIC	MEAN	ORGANIC	E MEAN	PARTICULATE SAMPLE	MEAN	PARTICULATE			

			/L) 0.0375 0.0290 0.003684	Mean Sediment (g/L) Sediment (e.d.) S.E.	S S &		0.0114 0.0068 0.00866	Mean Organics (g/L) Organics (s.d.) S.E.	8 Q E :				not evailable (*)	Data not avai
			0.0506				0.0100				0.0007	387.0	910531	15K24C
0.001426	0.0025	0.0539	0.0542	0.001870	0.0032	0.0132	0.0148	0,000901	0.0016	0.0670	0.0691	387 9	910531 910531	15K24A
			0.0658				0.0154				0.0812	374.2	910530	15K23C
	6.610	0.00	0.0006	0.00	0.00	0.0101	0.0147	6.00000	0.01.0	6,000	0.1053	374.2	910530	15K23B
0 00 7635	3	0799	0.0745		3		0.0183	200		0000	0.0028	350.9	910530	15K22C
			0.0812				0.0145				0.0957	350.0	910530	15K22B
0.007647	0.0132	0.0086	0.0503	0.001472	0.0026	0.0150	0.0122	0.008670	0.0150	0.0836	0.0625	350.0	910530	15K22A
			0.0059				0.0286				0.0945	322.7	910529	15K2 16
0.003315	0.0057	0.0714	0.0793	0.003824	0.0006	0.0221	0.0130	0.000512	0.0009	0.0935	0.0923	322.7	910529	15K21A
			4.643.4				• 60 10				0.0810	292.2	910529	15K20C
0.019777	0.0280	0.0380	0.0885	0.011509	0.0163	0.0326	0.0000	0.003814	0.0066	0.0758	0.0005	292.2	910529	15K20A
			•				•				0.0411	278.4	910529	15K 19C
•	•	•	••	•	•	•	• •	0.008359	0.0110	0.0566	0.0647	278.4 278.4	910529	15K 19A
,			0.0246				0.0097				0.0343	275.3	910528	15K 18C
0.012480	0.0210	0.0338	0.0000	0.002278	0.0039	0.0127	0.0183	0.014288	0.0247	0.0000	0.0045	275.3	910528	15K 18B
			0.0946				0.0168				0.1114	203.9	910528	15K 17C
6.00000		6.0064	0.0844	0.00000	0.0000	6	0.0180	0.000200			0.1024	203.0	910528	15K 17B
0 00000	200	9000	0.0351	000000		00179	0.0135		9	0007	0.0480	251.1	010528	15K 18C
			0.0617				0.0153				0.0770	251.1	910528	15K 168
0.006388	0.0111	0.0498	0.0832	0.001137	0.0020	0.0131	0.0140	0.000096	0.0118	0.0629	0.0072	243.8 251.1	910528 910528	15K 16A
			0.0912				0.0137				0.1049	243.0	910528	15K 15B
0.006634	0.0115	0.0794	0.0638	0.00 1033	0.0018	0.0151	0.0176	0.005033	0.0098	0.0945	0.0815	243.8	910528	15K 15A
			0.0544				0.0254				0.0798	225.4	910527	15K 14B
0.015713	0.0222	0.0267	0.0258	0.002007	0.0028	0.0225	0.0197	0.011152	0.0193	0.0758	0.0455	225.4	910527	15K 14A
			0.0439				0.0098	-			0.0535	201.2	910527	15K 13B
0.005361	0.0093	0.0387	0.0466	0.000554	0.0010	0.0090	0.0097	0.005914	0.0102	0.0477	0.0564	201.2	910527	15K 13A
			0.0357				0.0121				0.0478	200.0	910527	15K 12C
0.011244	0.0159	0.0134	0.0044	0.003792	0.0054	0.0046	0.0017	0.015036	0.0213	0.0180	0.0061	200.0	910527	15K 12A
			0.0292				0.0070				0.0302	175.4	910520	15K11C
0.006862	0.0110	0.0190	0.0255	0.002140	0.0037	0.0078	0.0124	0.008536	0.0148	0.0267	0.0380	175.4	910520	15K 1 1A
			0.0268				0.0117				0.0385	160.0	910520	15K 10C
0.00 1480	0.0020	0.0231	0.0213	0.001008	0.0027	0.0154	0.0105	0.000301	0.0000	0.0386	0.0378	500	910520	158 108
		}	0.0153				0.0084				0.0237	149.7	910520	15K9C
0.009615	0.0107	0.0384	0.0460	0.001425	0.0025	0.0117	0.0143	0.010900	0.0189	0.0501	0.0602	140.7	910526	15K9A
			0.0085				0.0005				0.0150	126.7	010525	15K8C
0.00	0.001	6.6140	0.0200	0.0000	9.00	0.0002	0.0041	0.002332	0.00	6.666	0.0240	126.7	010525	15K8B
0 00 27 10	0047	20145	0.0137	0 000047	3		0.0056	2000	3	0007	0.0192	124.0	910525	15K7C
			0.0173	!	!	!	0.0078	1			0.0251	124.0	910525	15K7B
808000	0 00 15	0 0155	0.0030	0 000521	0 00 00	0 0067	0.0045	0 00 1387	0 0004	0 0003	0.0075	100.0	910525	15K 6 C
			0.0112		i		0.0069		;	!	0.0181	100.0	910525	15K 0B
0.001976	0.0034	0.000	0.0062	0 00 18 19	0.0032	0 00 78	0.0030	0 002843	0 00 49	0 0 1 4 5	0.0092	77.0	910524	15K5C
!		!	•				•			!	•	77.0	910524	15K5B
0.000907	0.0013	0.0049	0.0036	0 000002	0.0000	0.0030	0.0030	0.002730	0.0030	0.0079	00000	77.0	810524	15K5A
							•				•	50.8	910524	15K4B
•	•	•	0.0104	•	•	0.0126	0.00126	0.006394	0.0000	0.0139	0.0230	50.0	910524	158.4
			0.0013				0.0085				0.0098	20.0	910523	15K38
0.000322	0.0006	0.0005	0.000	0.001788	0.0031	0.0042	0.0025	0.002098	0.0036	0.0047	0.0025	20.0	910523	15K3A
			0.0000				0.0073				0.0073	24.2	910523	15K2B
0.000252	0.0004	0.0006	0.0008	0.000913	0.0016	0.0051	0.0044	0.000068	0.0012	0.0057	0.0051	24.2	910623	15K2A
			0.0032				0.0021				0.0053	9 0	010523	15K 1B
0.000945	0.0013	0.0018	٠	0.000299	0,0004	0.0017	•	0.000047	0.0011	0.0038	0.0026	<u>0</u> .0	910523	15K 1A
!		9/1	9/L	!	•.d.	3/10	3/6	!	•	9/١	9/2	K	DATE	SAMPLE NO.
STANDARD	CONC	SEDIMENT	SEDIMENT	STANDARD	SAMPLE	ORGANIC	ORGANIC	STANDARD	CONC	PARTICULATE	PARTICULATE			
NEA X		MEAN		MEAN	ORGANIC	N A BA	(POM)	H MEAN	PARTICULAT	MEAN				

Table 3 - 1 Irradiametric values calculated at a discharge of $142 \text{ m}^3/\text{s}$.

910702	910702	910702	910701	910701	910630	910630	910630	910629	910629	910715	910715	910714	910714	910713	910713	910712	910712	910712	910628	DATE	
24	23	22	2																	SITE	
387.9	374.2	350.9	322.7	292.2	275.3	251.1	225.4	201.2	166.6	149.7	126.7	124.0	100.6	77.0	50.8	29.0	24.2	0.0	0.0	X S	
0.665	0.662	0.569	0.645	0.493	0.612	0.576	0.527	0.538	0.458	0.351	0.349	0.280	0.348	0.343	0.359	0.293	0.220	0.256	0.217	<u>ج</u>	
0.690	0.678	0.561	0.600	0.531	0.613	0.644	0.542	0.554	0.468	0.339	0.379	0.318	0.400	0.368	0.380	0.323	0.250	0.274	0.227	~	
0.697	0.688	0.561	0.598	0.538	0.608	0.665	0.543	0.550	0.467	0.341	0.389	0.324	0.421	0.373	0.384	0.323	0.255	0.272	0.232	7	
0.343	0.357	0.275	0.305	0.278	0.294	0.376	0.293	0.275	0.271	0.213	0.233	0.208	0.205	1.970	0.202	0.177	0.166	0.177	0.184	ω	
1637	2030	1407	1801	1687	1742	2044	1362	1599	1808	1817	1828	2199	1499	1928	1646	1682	2000	1610	1924	l(uE)	
6.01	6.37	6.77	6.35	8.18	6.64	7.34	7.25	7.39	8.95	11.68	11.79	15.33	11.25	12.13	11.17	13.73	19.06	15.54	19.17	OBS.	
6.31	6.35	7.39	6.51	8.52	6.87	7.30	7.98	7.80	9.17	11.96	12.05	14.99	12.08	12.23	11.71	14.33	19.06	16.39	19.35	CAL.	
4.62	4.10	3.68	3.71	3.47	4.24	3.32	3.32	4.00	3.12	1.87	2.02	1.40	1.70	1.94	2.51	5.33	0.82	1.36	NS S	Ţ	
0.0999	CBR	0.1039	CBR	0.1228	0.1228	0.1135	0.1135	0.1088	0.0890	0.0732	NAR	NAR	NAR	NAR	0.0737	NAR	NAR	NAR	NAR	Ra	
0.138	×	0.117	*	0.130	0.151	0.146	0.123	0.121	0.083	0.050	*	*	. *	*	0.056	*	×	×	*	م و	
0.041	*	0.034	*	0.036	0.042	0.041	0.035	0.035	0.025	0.016	*	*	*	×	0.018	*	*	*	*	5	
2.16	*	1.80	*	1.91	2.20	2.19	1.85	1.84	1.35	0.85	*	×	*	*	0.95	*	*	×	*	σ	
33.53	*	31.55	*	26.57	28.18	22.73	27.01	33.17	37.43	37.74	*	×	*	*	44.84	*	*	*	*	Tn/b' _b	
2.14	*	2.04	*	1.82	1.936	1.51	<u>-1</u> .8	2.18	2.31	2.21	*	×	×	×	2.63	×	*	×	*	Tn/b	

^{*} NS = No Sample; NAR = No Asymptotic Reflectance; and CBR = Channel Bed Reflectance

(Page 48)

9105	910530	910530	910529	910529	910528	910528	910527	910527	910526	910526	910525	910525	910525	910524	910524	910523	910523	910523	DATE	
																			• • •	
24	23	22	7	20	18	6	14	<u>1</u>	6	ဖ	œ	7	თ	Ŋ	4	ω	N	_	SITE	
387.9	374.2	350.9	322.7	292.2	275.3	251.1	225.4	201.2	166.6	149.7	126.7	124.0	100.6	77.0	50.8	29.0	24.2	0.0	X X	
1.033	0.927	0.923	1.169	0.735	1.093	1.085	0.893	1.077	0.706	0.689	0.465	0.392	0.351	0.389	0.299	0.283	0.242	0.239	ৣᠵ	
1.079	1.047	0.933	1.131	0.753	1.128	1.002	1.050	1.074	0.743	0.769	0.506	SN	0.368	0.398	0.315	S	0.247	0.230	مح	
S	S	SN	NS	S	S	SN	SN	S	S	S	NS.	SN	SN	S	SN	NS.	SN	NS S	7	
*	*	*	*	*	*	*	*	*	×	*	*	*	*	*	*	*	*	×	മ	
231	1839	1766	1536	1782	1658	1743	1532	1525	1534	1866	1861	2132	1679	489	1839	1466	1931	1919	l(uE)	
1.98	4.44	4.41	3.37	5.56	3.67	3.74	4.40	3.65	5.57	5.99	8.87	10.87	11.47	7.18	13.77	13.73	17.23	17.39	2	ᄁ
4.07	4.53	4.55	3.59	5.71	3.84	3.87	4.70	3.90	5.95	6.10	9.02	10.71	11.97	10.80	14.05	14.83	17.37	17.56	å Z	ALCUL
8.20	5.87	5.63	7.90	5.43	6.73	6.03	6.20	5.77	5.87	5.83	4.60	5.13	1.23	0.57	0.87	0.90	2.83	1.30	굽	ATED
CBR	0.1057	CBR	1.0520	CBR	0.0999	0.0999	CBR	0.0927	CBR	CBR	CBR	*	CBR	CBR	CBR	CBR	CBR	CBR	Ra	
*	0.221	*	0.238	*	0.225	0.200	*	0.199	*	*	*	*	*	*	*	*	*	*	م ق	
*	0.064	*	0.069	*	0.067	0.059	*	0.060	*	*	*	*	*	*	*	*	*	*	္မှာ	
*	3.40	*	3.66	*	3.53	3.13	*	3.19	*	*	*	*	*	*	*	×	*	*	σ	
*	26.51	*	33.20	*	29.87	30.13	*	28.96	*	*	*	*	*	*	*	*	*	*	Tn/b' _b	
*	1.73	*	2.16	×	1.91	1.93	*	1.81	*	×	*	×	*	*	*	*	*	*	Tn/b	

^{*} NS = No Sample; NAR = No Asymptotic Reflectance; and CBR = Channel Bed Reflectance

Table 5, Discharge 142 cms, geomorphological characteristics for channel topography are calculated from STARS Simulation Model data, correlated to irradiametric reach designation. These data are derived from bed material maps supplied by USGS. (Page 50)

24	23	22	21	20	18	16	14	13	10	9	8	7	თ	σı	4	ω	N		SITE
387.9	374.2	350.9	322.7	292.2	275.3	251.1	225.4	201.2	166.6	149.7	126.7	124.0	100.6	77.0	50.8	29.0	24.2	0.0	LOCATION KM
5.81	5.41	4.07	3.49	4.11	5.21	5.68	7.07	5.75	5.69	4.64	3.45	5.55	5.31	6.28	6.43	5.99	*	*	MEAN THALWEG DEPTH (M)
1.40	0.67	0.43	0.25	0.33	0.51	0.69	0.65	0.58	0.57	0.52	0.67	0.87	0.68	0.75	0.64	0.87	*	*	MEAN STANDARD ERROR (M)
225.85	188.77	236.54	174.10	165.80	134.70	158.24	153.26	142.17	185.51	204.01	293.55	269.06	260.41	185.09	159.05	284.69	*	*	TOP (M)
21.31	16.39	11.66	9.69	9.70	6.27	6.90	6.46	5.97	8.94	14.54	19.67	17.23	9.23	6.44	6.18	19.58	*	*	MEAN STANDARD ERROR (M)
0.00124	0.00127	0.00128	0.00137	0.00081	0.00112	0.00187	0.00156	0.00215	0.00190	0.00258	0.00198	0.00123	0.00078	0.00153	0.00144	0.00124	0.00124	*	SLOPE
0.665	0.662	0.569	0.645	0.493	0.612	0.576	0.527	0.538	0.458	0.351	0.349	0.280	0.348	0.343	0.359	0.293	0.220	0.217	8
<u>ი</u>	6.3	7.4	6.5	8.5	6.9	7.3	8.0	7.8	9.2	12.0	12.0	15.0	12.1	12.2	11.7	14.3	19.1	19.4	Zcp (M)

^{*} Missing values as indicated { * } are not avialable.

Mean geomorphic values for each irradiametric site are based on 1/2 the distance both upstream and downstream from the adjacent irradiametric sites of equal distances.

Table 6 — Discharge 425 cms, geomorphological characteristics for channel topography are calculated from STARS Simulation Model data, correlated to irradiametric reach designation. These data are derived from bed material maps supplied by USGS.

are calculated from (Page 51) are derived from

24	23	22	21	20	18	16	14	13	10	9	8	7	6	5 1	4	ယ	Ν	_		SITE	
387.9	374.2	350.9	322.7	292.2	275.3	251.1	225.4	201.2	166.6	149.7	126.7	124.0	100.6	77.0	50.8	29.0	24.2	0.0	, M	LOCATION	
6.98	7.24	5.84	5.27	6.26	7.17	7.63	8.68	7.76	7.20	6.18	4.93	5.77	6.68	7.93	8.34	7.65	*	*	<u>(</u>	DEPTH	MEAN THALWEG
1.43	0.72	0.48	0.28	0.34	0.55	0.72	0.68	0.60	0.59	0.57	0.70	0.88	0.71	0.77	0.66	0.88	*	*	(<u>M</u>	ERROR	MEAN STANDARD
236.50	205.62	258.77	205.07	195.25	153.29	179.51	175.87	162.26	201.99	227.93	296.19	304.23	278.40	200.16	174.42	302.33	*	*	<u>(</u>	WIDTH	TOP
20.69	15.61	11.38	10.30	9.48	5.76	6.82	6.45	5.61	8.75	15.51	20.28	17.85	10.44	6.45	6.60	18.67	*	*	(<u>X</u>	ERROR	MEAN STANDARD
0.00203	0.00141	0.00128	0.00141	0.00080	0.00117	0.00182	0.00164	0.00216	0.00190	0.00255	0.00198	0.00126	0.00082	0.00154	0.00144	0.00113	*	×			SLOPE
1.033	0.927	0.923	1.169	0.735	1.093	1.085	0.893	1.077	0.706	0.689	0.465	0.392	0.351	0.389	0.299	0.283	0.242	0.239		8	
4.1	4.5	4.6	3.6	5.7	3.8	3.9	4.7	3.9	5.9	6.1	9.0	10.7	12.0	10.8	14.0	14.8	17.4	17.6	<u>(</u> S	Zcp	

^{*} Missing values are indicated as { * }.

and downstream from the adjacent irradiametric sites. Mean geomorphic characteristics for each irradiametric site are based on 1/2 the distance both upstream

APPENDIX B

METHODS FOR DETERMINING PAR

(ATTENUATION, ABSORBANCE, SCATTERING, AND COMPENSATION DEPTHS)

AND

CORRELATION TECHNIQUES FOR OTHER METHODS OF LIGHT MEASUREMENT

METHODS FOR DETERMINING PAR:

Vertical Light Attenuation, (K)

One way to characterize optical properties of water is to determine the coefficient of extinction or attenuation (Wofsy 1983; Kirk 1977; Roemer and Hoagland 1979). The attenuation of *PAR* is described by the declining slope of subsurface irradiance, referred to as the attenuation coefficient, *K*. The attenuation coefficient can be calculated using a natural log regression of *PAR* measured at a series of depth intervals. This coefficient is considered one of the best descriptors for assessing the photosynthetic characteristics and capacities of water (Smith 1968; Kirk 1977; Kirk 1983).

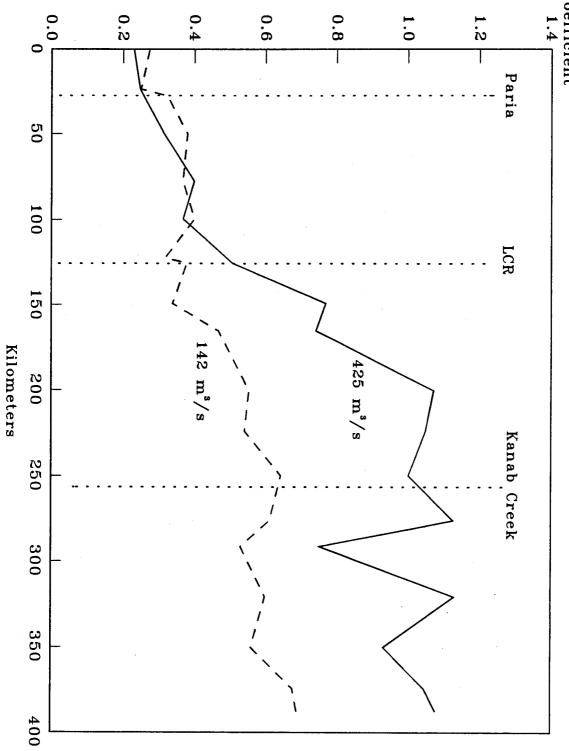
Subsurface intensities for all depths have been transformed to natural log values. The results of the linearization of subsurface intensities are exemplified in Fig. 11, illustrating progressive light attenuation for scalar irradiance at different sites located downstream from Glen Canyon Dam. Elevated attenuation coefficient values for K represent greater light attenuation. The differences in attenuation slope, K_{\circ} , between Site 2 (23.8 km), Site 9 (149.4 km) and Site 18 (276.3 km) demonstrates how light penetration is appreciably reduced downstream from Glen Canyon Dam.

Light Absorptance, (a)

As identified by Kirk (1977), a better understanding of the factors that contribute to light attenuation would be known if additional types of measurements were collected in conjunction with standard irradiametric measurements of total PAR. The absorptance coefficient, a, for a 400-700 nm bandwidth was determined in situ by a combination of measurements collected for downward, $E_{\rm d}$, and upward irradiance, $E_{\rm u}$, using cosine corrected sensors, and scalar irradiance, $E_{\rm o}$ using omni-directional sensors. This coefficient, a, represents the sum of the absorptive components consisting of water, soluble dyes, sediment, and organic particulates.

from Glen Canyon Dam (0 km) to Diamond Creek (387 km). at two steady state discharges of 142 m³/s and 425 m³/s on the Colorado River Figure 11 - Vertical attenuationcoefficients, (K_d), for cosine corrected measured





As suggested by Kirk (1983), the method for determining absorption coefficients a used in this study are based on the relationship expressed by eqn. 11.

$$a = K_{\rm E} E / E_{\rm o} \tag{11}$$

The absorption coefficient was determined using values for scalar irradiance, E_o , and net downward irradiance, E, consisting of the difference between downward irradiance, E_d , and upward reflectance, E_u , collected at a specific depth. The vertical attenuation coefficient, K_E , is derived from a natural log regression of net downward irradiance, E, for the entire depth profile.

Site differences for absorption coefficients were compared using a constant optical depth of 2.3 which represented the euphotic zones midpoint, $Z_{\rm m}$. The value 2.3 is the semi-log value of 10% of PAR. The absorption coefficient for a given depth differs slightly due to water surface variability and channel reflectance. The selection of $Z_{\rm m}$ compensated for the variability in absorbance observed throughout the depth profile. The calculated absorption coefficient, a, at $Z_{\rm m}$, and the measured absorptance range for the entire depth profile are depicted in Fig 7. The in situ measurement represents the total absorption spectra including the absorptive constituents of water, dissolved and particulate color, and particulates of inorganic and organic nature.

Light Scattering, (b)

Total scattering and normal backscattering coefficients are very difficult to measure either *in situ* or laboratory settings. An alternative approach has been suggested by Kirk (1980a) for calculating different scattering coefficients. Irradiametric data were used to determine the asymptotic backscattering coefficient, b'_b . A relationship exists for asymptotic reflectance and scattering, founded on the theory that any horizontal or vertical reflectance in water is a result of a backscattering (90-180°) phenomena (Kirk 1977, Kirk 1980a). Asymptotic backscattering

coefficient, b'_b , defined as the portion of downward irradiance scattered backwards (180°) is easily measured in the field using the collected irradiametric data at the depth where the ratio of upward reflectance, E_u , to downward irradiance, E_d , quanta irradiance reaches a point of equilibrium.

The asymptotic reflectance, $R_{\rm a}$, is determined by assuming that the attenuation coefficient, K, does not change with depth (Kirk 1983). The depth at which asymptotic reflectance, $R_{\rm a}$, occurs in the Colorado River was considered the leveling off point of the reflectance ratio, $E_{\rm u}$ / $E_{\rm d}$ (Di Toro 1978; Kirk 1977; and Kirk 1980). A multiple line regression was used to determine the line intercept. It was possible to calculate the asymptotic backscattering coefficient, $b'_{\rm b}$, by using eqn. 12 (Kirk 1980a).

$$b'_{b} = 2 \cdot R_{a} \cdot K \tag{12}$$

Though important, the conversion of the measured asymptotic backscattering coefficient, b_b' , to a more useable form such as normal backscattering coefficient, b_b , or total scattering coefficient, b_b , provides more insightful information on the actual scattering properties of water. Normal backscattering, b_b , is determined using the data assembled by Kirk (1980a) for irradiant distributional data of varying turbidities. A linear relationship has been established for known values of asymptotic reflectance, R_a , to ratios of asymptotic backscattering coefficient and normal backscattering coefficients, b'_b/b_b , for varying levels of turbidity. The regression equation (eqn. 13) used for determining the normal backscattering coefficients.

$$b'_{b}/b_{b} = 10.468 \cdot R_{a} + 2.342, R^{2} = 0.984$$
 (13)

Normal backscattering coefficient, $b_{\rm b}$, are derived by dividing the conversion ratio $b'_{\rm b}$ / $b_{\rm b}$ into the calculated value for asymptotic backscattering, $b'_{\rm b}$. Total scattering coefficient, b, can be determined using the mean value for the normal backscattering coefficient of 0.0190 m⁻¹ (Petzold 1972) relative to the total scattering coefficient of 1.0 m⁻¹ (Kirk 1980). The total scattering coefficient, b, equals the value of $b_{\rm b}$ multiplied by 53, a factor derived from the ratio of $b_{\rm b}$ to b. This relationship was developed through Petzold's research (1972) on scattering functions for oceanic waters. Additional information specific to the methods for deriving different types of scattering coefficients, $b'_{\rm b}$, $b_{\rm b}$, and b and their relationships are treated in great detail by both Kirk's (1977; 1980a; and 1983) and Petzold's (1972) research.

Compensation Depth, Z_{cp}:

The euphotic zone depth, $Z_{\rm eu}$, defined as 1% of PAR, has been used by researchers for estimating the extent of the underwater light regime. Talling (1965), identified a relationship between the optical depth of 3.7 and the measured attenuation coefficient K for East African lakes where light attenuation was attributed to high densities of phytoplankton (Ganf 1974). The use of this value and the inverse relationship between euphotic zone, $Z_{\rm au}$, and the vertical attenuation coefficient, K, have been verified in other lakes of similar optical conditions. Kirk (1983) identified the use of 4.6 as a constant representing the natural log relationship of 1% PAR. This logrythmic relationship provides a sufficient estimator of the euphotic zone depth for waters similar to those of the Colorado River, expressed as $Z_{\rm eu} = 4.6/K$. However, this relationship does not take into account vertical movement of the euphotic zone due to incidental light variation encountered during seasonal and diel shift of the photoperiod or atmospheric conditions. The optical depth relationship of 4.6/K was not considered as reliable a method for determining actual or calculated compensation points because of the possibility of either overestimating or underestimating the actual depth.

As an alternative method we suggest using an equation for calculating depths for the compensation

point $Z_{\rm cp}$. The natural log value for surface intensity, I, subsurface intensity, $I_{\rm ss}$, and the attenuation coefficient, K, are used for determining actual depths and calculated depths at variable surface intensities. These depths were calculated by solving for the line-intercept using the attenuation slope (K) and the mean incidental light intensity measured during the depth profile. The compensation point depth equation used is expressed in eqn. 14.

$$Z_{\rm cp} = -1/K \cdot \ln I + 1/K \cdot \ln I_{\rm ss}. \tag{14}$$

The natural log values for surface intensity, I, subsurface intensity, I_{ss} , and the attenuation coefficient, K, are used for determining actual depths and calculated depths in the equation. The value 3.4 is a natural log transformation of 30 μ E which represents the compensation point for C. glomerata (Graham 1982). This constant (3.4) specific to C, glomerata or other aquatic algae sharing similar compensation points and is expressed in eqn. 15, having been derived from the above equation.

$$Z_{cp} = 3.4 - \ln l / K \tag{15}$$

The calculated compensation depths, $Z_{\rm cp}$, at a constant 2000 $\mu{\rm E}$ surface intensity allow for a depth comparison between sites and discharges by controlling the incidental light source. We were unable to collect a representative sample of subsurface intensities at all sites; therefore, subsurface intensities were not used because consistency in data collection were often hampered by a wide range of incidental solar flux or quanta sensor movement within the air-water interface. Due to problems encountered in measuring subsurface intensities, $I_{\rm ss}$, in a riverine environment, we elected to replace actual subsurface values with incidental light intensities, I. For comparative reasons, incidental surface intensities were adjusted to 2000 $\mu{\rm E}$ and used with the attenuation coefficient, K, for calculating the compensation depth rather then the actual *in situ* compensation depth.

It was found that the calculated compensation depth underestimates the compensation depth for scalar irradiance using surface values rather than subsurface values. However, measurements based on incidental surface readings are more consistent and allow for comparisons between sites. The mean difference in calculated depth are $0.28 \text{ m} \pm 0.16 \text{ (s.d.)}$, based on subsurface measurements to incidental light measurements.

CORRELATIONS BETWEEN SCALAR IRRADIANCE AND LIGHT MEASUREMENT METHODS Correlations for Downward Irradiance

Vertical attenuation coefficients for scalar irradiance, K_o , and downward (cosine corrected) irradiance, K_d , of PAR were calculated for each sample site and replicated at each of the different research flows (142 and 425 m³/s). Attenuation coefficients for scalar irradiance, K_o , and cosine corrected irradiance, K_d , are graphically depicted (Fig 12 and 13) and the tabulated results for each site are in Tables 3 and 4. Data analysis specific to each site and discharge are found in Appendix D.

Vertical attenuation coefficients derived from measurements of scalar irradiance and downward irradiance are significantly different (paired-t = -3.346, df = 1,36 , p = 0.00193). In the Colorado River under identical optical light conditions total PAR will be slightly underestimated using attenuation coefficients derived from downward irradiance, K_d . However, using a regression equation downward irradiance, K_d , can be made equivalent ($F_{1,36}$ = 1719.3 ; p < 0.0001) to scalar irradiance, K_o . This equation (eqn. 16) will allow researchers to interrelate data collected using different sensor types for measuring quanta irradiance.

$$K_o \equiv 0.978 \cdot K_d - 0.00957 \ (R_{adj}^2 = 0.979)$$
 (16)

Correlations for Secchi Depth

Secchi depth measurements are a standard limnological tool for measuring depths of light reflectance. A method has been developed from our analysis to determine vertical attenuation coefficients from secchi depths collected in the Colorado River. Based on the degree of water clarity two separate regressions were developed for correlating secchi depth measurements, Z_{SD} , to attenuation coefficients, K_o , for scalar irradiance. The selection of the proper equation is dependent on the maximum secchi depth measurement; for depths measured ≥ 2.7 m select eqn. 17;

$$K_{\circ} = -0.243 \cdot \ln Z_{SD} + 0.711, (R^2 = 0.892)$$
 (17)

and for depths ≤ 2.7 m, select eqn. 18,

$$K_{\circ} = -0.857 \cdot \ln Z_{SD} + 1.243, (R^2 = 0.868)$$
 (18)

A relationship of secchi depths in excess of 8 m to light attenuation has not been effectively developed at this point in time.

APPENDIX C

MANAGEMENT CONSIDERATIONS FOR EIS-ALTERNATIVES:

MANAGEMENT CONSIDERATIONS FOR EIS-ALTERNATIVES:

The proposed alternatives for the operations at Glen Canyon Dam are sure to affect the availability of light for photosynthesis in the Colorado River. The optical properties of water are functionally controlled by the attenuating components which include water, soluble dyes, and suspended organic and particulate organic material. At present the primary light attenuating component in the Colorado River is suspended sediment. The complex interaction between the hydrology, geomorphology, and sediment contribution influence the processes of sediment storage, transport, degradation and aggradation in the riverine system. In regards to the operations of Glen Canyon Dam, certain EIS-alternatives pose both negative and positive affects to light availability and the potential photosynthetic productivity. An understanding of the factors which control primary production during periods of optimum and semi-optimum light conditions need to be further investigated. Discussed below, is a review of each EIS-alternative as it applies to photosynthetically available light.

Year Round Steady Flow

Low steady flows (142 m³/s) have been observed to decrease sediment transport capacity which would result in lower light attenuation (Cluer 1992), (refer to Table 1, and Fig. 3). These types of flows would potentially reduce transport of suspended sediment and increase available light except for periods of sediment discharge from tributaries above base flow. However, as identified in Fig. 9, a reduction in submerged area available for photosynthesis occurs due to changes in the vertical stage discharge. Discharge volume adjusts to the channel topography as mean channel depth and top width shift in response to changes in vertical stage (Randel and Pemberton 1987).

High steady flows (425 m³/s) appear to increase the transport and suspension of alluvial sediment stored in the river channel and banks and thereby increasing light attenuation, (refer to Table 2, and Fig. 4). Secondly, a cascading affect results in sediment remaining in suspension into downstream

areas hydrologically similar to areas upstream having low light attenuation (Fig. 1.5). It is speculated that during periods of steady flood flows comparable to 1983-86, would further amplify the light attenuating process in the river. As identified by Pemberton (1987), measurements for suspended sediment concentrations increased during these flood periods. High steady flows would create semi-optimum light conditions for photosynthesis by reducing the compensation point for <u>C. glomerata</u> above a 4 m depth. Partial light extinction would further reduce available area for colonization and algal growth. The affect on production would be more restrictive from the Granite Gorge (150 km) to Diamond Creek (387 km), (refer to Fig. 10).

Seasonally Adjusted Steady Flows

Seasonally adjusted steady flows have the potential to optimize light availability by seasonally adjusting discharge volume with regard to sediment discharge from the primary tributaries (e.g. Paria, LCR, and Kanab Creek). However, higher flows during high delivery months (April - June) would increase light attenuation through the transport and storage of alluvial sediment in the river channel than during low steady flows that occur during months of high tributary runoff (Herford 1984; and Graf *et al.* 1991).

Existing Monthly Volume Steady Flows

As with seasonally adjusted steady flows light availability would increase during low steady flows but would be decreased during higher water delivery months. These critical delivery periods generally coincide during the same time periods when there are minimal input of sediment from tributaries. At these time periods there is a high percentage of available days for optimum light conditions. The effect from this flow alternative is hard to predict since little is known about either the fluvial dynamics or resident time of sediment storage in the channel bed under various flow scenarios. Until such an understanding exists, it will be difficult predicting the sediment available for transport and the corresponding attenuation of light in the water column.

Low Fluctuating Flows

As discussed above, Low Fluctuating Flows would probably result in higher light attenuation than a comparable discharge volume under steady flow conditions. However, the degree of light attenuation increase is unknown and is dependent on the stability of alluvial and channel deposits.

Moderate Fluctuating Flows

Moderate Fluctuating Flows would probably further increase light attenuation than low fluctuating flows. The response time and residency of sediment in suspension with the declining limb of the hydrograph are unknown. As identified in Fig. 4, the reduction in suspended sediment concentration is not instantaneous in geographical reaches where there are distinct hydrological changes.

High Fluctuating Flows

High fluctuating flows are expected to have the same effect on the availability of light as the No Action Alternative.

No Action Alternative

During periods of tributary base flow, the availability of light in irradiametric reaches below the confluence of the Little Colorado River will remain for large periods of time under sub-optimum to light extinction conditions. Secondly, the diel pattern of discharge releases will become synchronized with the photoperiod sustaining certain portions of the river corridor under a particular discharge regimen, (i.e., high-descending and low ascending). The exception to this would be the weekly disruption in discharge volume as week-day releases shift into week-end releases.

ADDITIONAL ELEMENTS TO FLOW ALTERNATIVES

Beach Protection

This would probably have a positive affect on the amount of light available for photosynthesis by stabilizing alluvial sediments that would otherwise be re-suspended during degradational flow patterns.

Beach/Habitat Building Flows

These flows would have negative impacts to availability of light during the actual flow event. The quasi-equilibration rate of alluvial deposition after these flows are unknown. If these flows are implemented they should be timed to coincide with sediment discharge from downstream tributaries to avoid either additional light attenuation or extinction which normally results during these events.

Sediment Augmentation

Sediment augmentation would have a negative impact on the availability of light for photosynthesis. If sediment augmentation is implemented it should be timed to coincide with high sediment discharge from tributaries which would otherwise cause light extinction. It is speculated, that large inputs of sediment into the river system would readjust the sediment storage in the river channel; this additional sediment would be potentially available for transport. This would probably result in increasing light attenuation. Differences in light attenuation and storage of alluvial sediment were observed when comparing dissimilarities above and below both the Paria and the Little Colorado River tributaries.

Multi-level Withdrawal Intake Structure

The optical properties found in the river are affected by the light attenuating constituents in the water source. Increases in turbidity from suspended inorganic and organic material would further

attenuate the optical properties of the river. These attenuating changes have been observed, but the degree of maximum light attenuation are difficult to assess.

APPENDIX D

REGRESSION ANALYSIS FOR STUDY SITES

MEASURED AT DISCHARGES OF 142 AND 425 m³/s

Ko Coefficient(s) 0.241757 Std Err of Coef. 0.002382	Regression Output: 7.566477 Constant 7.566477 Std Err of Y Est 0.044145 R Squared 0.996225 No. of Observations 39	Site 2, Date 910523, Time: (1348–1421) Discharge 425 m³/s Location: 23.8 km	Ko Coefficient(s) 0.302935 Std Err of Coef. 0.003789	Regression Output 5.774843 Constant 5.774843 Std Err of Y Est 0.144084 R Squared 0.955023 No. of Observations Degrees of Freedom 301	Site 1, Date 911216, Time: (1239–1317) Discharge 376–396 m³/s Location: 0.0 km	Ko Coefficient(s) 0.239180 Std Err of Coef. 0.020395	Regression Output: 7.447537 Constant 7.447537 Std Err of Y Est 0.315995 R Squared 0.741284 No. of Observations 50 Degrees of Freedom 48	Site 1, Date 910523, Time: (1053–1119) Discharge 425 m³/s Location: 0.0 km
	Terrestrial Intensity Minimum 1854 Maximum 2004 Mean 1930.756 Std 25.814			Terrestrial Intensity Minimum 245.1 Maximum 386.6 Mean 289.4864 Std. 32.388			Terrestrial Intensity Minimum 1851 Maximum 1978 Mean 1919.2 Std. 34.857	
Ko Coefficient(s) 0.220369 Std Err of Coef. 0.001817	Regression Output: 7.579088 Constant 7.579088 Std Err of Y Est 0.041991 R Squared 0.995933 No. of Observations 62 Degrees of Freedom 60	Site 2, Date 910712, Time: (1127–1213) Discharge 142 m³/s Location: 23.8 km	Ko Coefficient(s) 0.256243 Std Err of Coef. 0.002328	Regression Output: Constant 7.444679 Std Err of Y Est 0.073927 R Squared 0.995151 No. of Observations 61 Degrees of Freedom 59	Site 1, Date 910712, Time: (0933–1008) Discharge 142 m³/s Location: 0.0 km	Ko Coefficient(s) 0.217134 Std Err of Coef. 0.003779	Regression Output: Constant 7.469774 Std Err of Y Est 0.106654 R Squared 0.981853 No. of Observations 63 Degrees of Freedom 61	Site 1, Date 910628, Time: (1339–1419) Discharge 142 m³/s Location: 0.0 km
	Terrestrial Intensity Minimum 1946 Maximum 2032 Mean 1999.548 Std. 19.797			Terrestrial Intensity Minimum 1521 Maximum 1672 Mean 1609.557 Std. 44.992			Terrestrial Intensity Minimum 1853 Maximum 2003 Mean 1923.571 Std. 45.21	

Site 2, Date 910711, Time: (1334-1413) Discharge 759-770 m³/s

Location: 23.8 km

Degrees of Freedom	No. of Observations				Regression Output:
44	46	0.990049	0.052339	7.666379	
	Std.	Mean	Maximum	Minimum	Terrestrial
	30.225	1946.675	2006	Ainimum 1906	Intensity

Std Err of Coef. Ko Coefficient(s) 0.004564 0.30203

Site 3, Date 910523, Time: (1528-1556)

Discharge 425 m³/s Location: 27.74 km

R Squared Std Err of Y Est Constant Regression Output:

Ko Coefficient(s)
Std Err of Coef. Degrees of Freedom No. of Observations 0.007353 0.28326 0.123147 0.972477 7.340761 Mean Maximum Minimum Terrestrial Intensity 1466.113 39.842 1410 1560

Site 4, Date 910524, Time: (1016-1052)

Location: 50 km Discharge 425 m³/s

R Squared Std Err of Y Est Degrees of Freedom No. of Observations Constant Regression Output 0.997094 0.054133 7.572673

Mean Std.

1839.259

37.66 1890

Minimum

1747

Terrestrial Intensity

Maximum

Ko Coefficient(s) Std Err of Coef. 0.001862 0.298870

Site 3, Date 910712, Time: (1433-1524)

Discharge 142 m³/s Location: 27.74 km

Std Err of Y Est Degrees of Freedom No. of Observations R Squared Constant Regression Output: 0.987876 0.0658587.537181 Mean Std. Maximum Minimum Terrestrial Intensity

1681.9 44.212

1773 1624

Ko Coefficient(s) Std Err of Coef. 0.293177 0.004503

Discharge 142 m³/s Site 4, Date 910713, Time: (0928-1030)

Location: 50 km

Degrees of Freedom No. of Observations R Squared Constant Std Err of Y Est Regression Output: 0.990570 0.087489 7.559331

Mean Std. Maximum Minimum Terrestrial Intensity 1646.187 44.212 1566 1715

Ko Coefficient(s) Std Err of Coef. 0.358679 0.004444

Regression Output: Constant 7.850438 Std Err of Y Est 0.094655 R Squared 0.991608 No. of Observations 0.991608 Coefficient(s) 0.392197 Std Err of Coef. 0.004509	Site 7, Date 910525, Time: (1311–1345) Discharge 425 m³/s Location: 122.83 km	Regression Output: Constant Constant 7.590677 Std Err of Y Est No. of Observations Degrees of Freedom Ko Coefficient(s) Coeff	Site 6, Date 910525, Time: (0932–1000) Discharge 425 m³/s Location: 99.5 km	Regression Output: Constant 6.360632 Std Err of Y Est 0.126415 R Squared 0.977043 No. of Observations 11 Degrees of Freedom 9 Ko Coefficient(s) 0.388773 Std Err of Coef. 0.019864	Site 5, Date 910524, Time: (1512–1535) Discharge 425 m³/s Location: 77.61 km
Terrestrial Intensity Minimum 2052 Maximum 2198 Mean 2131.696 Std. 34.036		Terrestrial Intensity Minimum 1604 Maximum 1732 Mean 1678.528 Std. 24.78		Terrestrial Intensity Minimum 403 Maximum 544 Mean 489.427 Std. 48.571	
Regression Output: Constant 7.701544 Std Err of Y Est 0.045699 R Squared 0.996210 No. of Observations Degrees of Freedom 67 Ko Coefficient(s) Std Err of Coef. 0.002111	Site 7, Date 910714, Time: (1245–1321) Discharge 142 m³/s Location: 122.83 km	Regression Output: Constant 7.538938 Std Err of Y Est 0.056397 R Squared 0.991742 No. of Observations Degrees of Freedom 47 Ko Coefficient(s) 5td Err of Coef. 0.004730	Site 6, Date 910714, Time: (0921–1005) Discharge 142 m³/s Location: 99.5 km	Regression Output: Constant 7.719158 Std Err of Y Est 0.093288 R Squared No. of Observations Degrees of Freedom 92 Ko Coefficient(s) Std Err of Coef. 0.002502	Site 5, Date 910713, Time: (1325–1413) Discharge 142 m³/s Location: 77.61 km
Terrestrial Intensity 1544 Minimum 2122 5699 Maximum 2270 6210 Mean 2199 69 Std. 42.712	321)	Terrestrial Intensity 8938 Minimum 1430 6397 Maximum 1573 1742 Mean 1498.5 47 Std. 42.601 45	005)	Terrestrial Intensity 9158 Minimum 1876 3288 Maximum 1974 5136 Mean 1928.2 94 Std. 27.403	413)

Location: 128.56 km Discharge 425 m³/s Site 8, Date 910525, Time: (1423-1459) Discharge 142 m³/s

Constant
Std Err of Y Est R Squared No. of Observations Regression Output: 7.821427 0.091055 0.994377 Mean Maximum Minimum Terrestrial Intensity 1860.9 40.088 1939 1801

Std Err of Coef. Ko Coefficient(s) 0.465453 0.006614 Degrees of Freedom

Site 9, Date 910526, Time: (1032-1107)

Location: 149.38 km Discharge 425 m³/s

R Squared Std Err of Y Est Std Err of Coef. Ko Coefficient(s) Constant Degrees of Freedom No. of Observations Regression Output: 0.688990 0.005100 0.125566 0.996179 7.833957 Mean Maximum Minimum Terrestrial Intensity 1866.305 37.475 1920 1787

Site 10, Date 910526, Time: (1511-1542) Discharge 425 m³/s

Location: 165.47 km

R Squared Std Err of Y Est No. of Observations Constant Regression Output: 0.994086 0.188447 7.714170 Mean Std. Maximum Minimum Terrestrial Intensity

1534.236 47.828

1600

1450

Ko Coefficient(s) Std Err of Coef. 0.706328 0.006511

Degrees of Freedom

Site 8, Date 910715, Time: (1415-1533)

Location: 128.56 km

Degrees of Freedom No. of Observations R Squared Constant Std Err of Y Est Regression Output: 0.038054 0.9982457.756251

Mean

1828.1

1893

50.256

Minimum

Terrestrial Intensity

Maximum

Std Err of Coef. Ko Coefficient(s) 0.348634 0.002109

Site 9, Date 910715, Time: (1042-1137)

Location: 149.38 km Discharge 142 m³/s

Degrees of Freedom R Squared Std Err of Y Est Constant No. of Observations Regression Output: 7.741661 0.996959 0.100084 Mean Maximum Minimum Terrestrial Intensity

1816.7

1900

1751

43.878

Std Err of Coef. Ko Coefficient(s) 0.351247 0.002270

Site 10, Date 910629, Time: (1011-1050)

Discharge 142 m³/s Location: 165.47 km

R Squared Std Err of Y Est Degrees of Freedom No. of Observations Constant Regression Output: 0.075293 0.997454 7.655929

> Mean Maximum Minimum

74.673 1808.1

1906 1758

Terrestrial Intensity

Ko Coefficient(s) Std Err of Coef. 0.002524 0.458096

Ko Coefficient(s) 1.085479 Std Err of Coef. 0.012658	Regression Output: Constant 7.947494 Std Err of Y Est 0.162688 R Squared 0.993380 No. of Observations 51 Degrees of Freedom 49	Site 16, Date 910528, Time: (1005–1040) Discharge 425 m³/s Location: 249.94 km	Ko Coefficient(s) 0.893091 Std Err of Coef. 0.014424	Regression Output: Constant 7.739329 Std Err of Y Est 0.269485 R Squared 0.984589 No. of Observations 62 Degrees of Freedom 60	Site 14, Date 910527, Time: (1515–1549) Discharge 425 m³/s Location: 224.03 km	Ko Coefficient(s) 1.07666 Std Err of Coef. 0.021384	Regression Output: Constant 8.265475 Std Err of Y Est 0.359879 R Squared 0.981036 No. of Observations 51 Degrees of Freedom 49	Site 13, Date 910527, Time: (1116–1140) Discharge 425 m³/s Location: 200.38 km
	Terrestrial Intensity Minimum 1684 Maximum 1809 Mean 1742.607 Std. 37.741			Terrestrial Intensity Minimum 1462 Maximum 1612 Mean 1531.854 Std. 53.107			Terrestrial Intensity Minimum 1952 Maximum 2019 Mean 1983 Std. 15.3797	
Ko Coefficient(s) 0.575515 Std Err of Coef. 0.006384	Regression Output: Constant 7.938944 Std Err of Y Est 0.083529 R Squared 0.994006 No. of Observations 51 Degrees of Freedom 49	Site 16, Date 910630, Time: (1206–1225) Discharge 142 m³/s Location: 249.94 km	Ko Coefficient(s) 0.526510 Std Err of Coef. 0.004246	Regression Output: Constant 7.599766 Std Err of Y Est 0.082871 R Squared 0.996434 No. of Observations 57 Degrees of Freedom 55	Site 14, Date 910630, Time: (0856–0925) Discharge 142 m³/s Location: 224.03 km	Ko Coefficient(s) 0.538365 Std Err of Coef. 0.005038	Regression Output: Constant 7.561004 Std Err of Y Est 0.133153 R Squared 0.994166 No. of Observations 69 Degrees of Freedom 67	Site 13, Date 910629, Time: (1525–1603) Discharge 142 m³/s Location: 200.38 km
	Terrestrial Intensity Minimum 2026 Maximum 2050 Mean 2043.6 Std. 5.365	•		Terrestrial Intensity Minimum 1331 Maximum 1481 Mean 1361.6 Std. 47.582			Terrestrial Intensity Minimum 1519 Maximum 1667 Mean 1597.8 Std. 44.825	

Regression Output: Constant 8.033757 Std Err of Y Est 0.363721 R Squared 0.988901 No. of Observations Degrees of Freedom 50	Site 21, Date 910529, Time: (1503–1523) Discharge 425 m³/s Location: 321.06 km	Ko Coefficient(s) 0.734972 Std Err of Coef. 0.007656	Regression Output: 7.908364 Constant 7.908364 Std Err of Y Est 0.101961 R Squared 0.994603 No. of Observations 52 Degrees of Freedom 50	Site 20, Date 910529, Time: (1036-1056) Discharge 425 m³/s Location: 291.61 km	Ko Coefficient(s) 1.092970 Std Err of Coef. 0.011555	Regression Output: 7.936153 Constant 7.936153 Std Err of Y Est 0.191473 R Squared 0.99358 No. of Observations 60 Degrees of Freedom 58	Site 18, Date 910528, Time: (1439–1456) Discharge 425 m³/s Location: 276.33 km
Terrestrial Intensity Minimum 144 Maximum 159 Mean 1535.96 Std. 44.63			Terrestrial Intensity Minimum 174 Maximum 181 Mean 1782.09 Std. 23.459			Terrestrial Intensity Minimum 161 Maximum 170 Mean 1657.1 Std. 28.3	
ntensity 1448 1591 1535.961 44.632			ntensity 1744 1819 1782.096 23.4596			ntensity 1612 1702 1657.55 28.353	
Regression Output: Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom	Site 21, Date 910701, Time: (1425–1506) Discharge 142 m³/s Location: 321.06 km	Ko Coefficient(s) 0.492852 Std Err of Coef. 0.003695	Regression Output: Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom	Site 20, Date 910701, Time: (Discharge 142 m³/s Location: 291.61 km	Ko Coefficient(s) 0.611572 Std Err of Coef. 0.003222	Regression Output: Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom	Site 18, Date 910630, Time: (1447–1513) Discharge 142 m³/s Location: 276.33 km
: 7.886510 0.935055 0.861380 79	14251506)		7.753634 0.058755 0.996861 58)1, Time: (0953–1023) ; n		7.851642 0.062800 0.998281 64 62	1447–1513)
Terrestrial Intensity Minimum 172 Maximum 187 Maximum 1801 Mean 43.13			Terrestrial Intensity Minimum 162 Maximum 175 Mean 1686 Std. 41.41			Terrestrial Intensity Minimum 167 Maximum 180 Mean 1741 Std. 32.64	
ntensity 1728 1873 1801.1 43.137			ntensity 1623 1753 1686.5 41.418			ntensity 1672 1808 1741.7 32.645	

Ko Coefficient(s) Std Err of Coef.

1.168774 0.017510

Ko Coefficient(s) Std Err of Coef.

0.644801 0.029477

Ko Coefficient(s) 1.033104 Std Err of Coef. 0.017144	Regression Output: Constant 5.893988 Std Err of Y Est 0.198175 R Squared 0.989643 No. of Observations Degrees of Freedom 38	Site 24, Date 910531 , Time: (1000–1020) Discharge 425 m³/s Location: 387.51 km	Ko Coefficient(s) 0.926508 Std Err of Coef. 0.007439	Regression Output: Constant 7.91 0833 Std Err of Y Est 0.111755 R Squared 0.997171 No. of Observations Degrees of Freedom 44	Site 23, Date 910530, Time: (1411–1501) Discharge 425 m³/s Location: 374.15 km	Ko Coefficient(s) 0.923394 Std Err of Coef. 0.004846	Regression Output: 7.843190 Constant 7.843190 Std Err of Y Est 0.097131 R Squared 0.998185 No. of Observations Degrees of Freedom 66	Site 22, Date 910530, Time: (1012–1039) Discharge 425 m³/s Location: 350.18 km
	Terrestrial Intensity 988 Minimum 223.2 175 Maximum 361.7 643 Mean 230.7 40 Std. 36.628	020)	\$ 50 S	Terrestrial Intensity 833 Minimum 1742 755 Maximum 1899 171 Mean 1839.26 46 Std. 34.621	501)		Terrestrial Intensity 190 Minimum 1707 131 Maximum 1816 185 Mean 1765.588 68 Std. 31.795 66	039)
Ko Coefficient(s) 0.665269 Std Err of Coef. 0.003842	Regression Output: Constant 7.812365 Std Err of Y Est 0.062675 R Squared 0.998102 No. of Observations Degrees of Freedom 57	Site 24, Date 910702, Time: (1513–1541) Discharge 142 m³/s Location: 387.51 km	Ko Coefficient(s) 0.66145 Std Err of Coef. 0.004957	Regression Output: Constant 7.992922 Std Err of Y Est 0.085299 R Squared 0.996696 No. of Observations Degrees of Freedom 59	Site 23, Date 910702, Time: (1314–1348) Discharge 142 m³/s Location: 374.15 km	Ko Coefficient(s) 0.568634 Std Err of Coef. 0.003121	Regression Output: Constant 7.611927 Std Err of Y Est 0.086006 R Squared 0.998075 No. of Observations Degrees of Freedom 64	Site 22, Date 910702, Time: (08520934) Discharge 142 m³/s Location: 350.18 km
	Terrestrial Intensity Minimum 1553 Maximum 1703 Mean 1637.1 Std. 48.546		***************************************	Terrestrial Intensity Minimum 2003 Maximum 2082 Mean 2030 Std. 17.363		巻 (Terrestrial Intensity Minimum 1333 Maximum 1482 Mean 1407.4 Std. 40.076	

Site 1, Date 910523, Time: (1129-1215) Discharge 425 m³/s

Location: 0.0 km

Kd Coefficient(s) 0.2 Std Err of Coef. 0.0	Degrees of Freedom	No. of Observations	R Squared	Std Err of Y Est	Constant	Regression Output
0.230401 0.005205	74	76	0.963605	0.151868	7.430618	Aput .
		Std.	Mean	Maximum	Minimum	Terrestrial Intensity
		25.218	2021.934	2068	1964	Intensity

Site 1, Date 910628, Time: (1339-1458) Discharge 142 m³/s R Squared Std Err of Y Est Constant Location: 0.0 km Regression Output: 0.980678 0.114991 7,333657

Kd Coefficient(s) Std Err of Coef. 0.004085

Degrees of Freedom No. of Observations

ණු සු

Mean Std. Maximum

1918.158

1989

1848

43.959

Minimum

Terrestrial Intensity

Discharge 142 m³/s

Site 1, Date 910712, Time: (0925-1018)

Location: 0.0 km

Site 1, Date 911216, Time: (1239–1317) Discharge 376–396 m³/s

Location: 0.0 km

Degrees of Freedom	No. of Observations	R Squared	Std Err of Y Est	Constant	Regression Output:
328	330	0.965903	0.149289	5.560097	

Mean Maximum Minimum Terrestrial Intensity 34.40311 289.6415 386.6

Degrees of Freedom Std Err of Y Est Std Err of Coef. Kd Coefficient(s) No. of Observations R Squared Constant Regression Output: 0.273746 0.003340 0.992021 0.102471 7.200940

> Mean Maximum Minimum Terrestrial Intensity 1593.553 43.606 1505 1651

Site 2, Date 910523, Time: (1300-1337) Discharge 425 m³/s

Kd Coefficient(s) Std Err of Coef.

0.33675 0.89334

Location: 23.8 km

0.091952 0.982008 7,400079 Mean Maximum Minimum Terrestrial Intensity 2017.375 16.68 2049 1969

> Location: 23.8 km Discharge 142 m³/s Site 2, Date 910712, Time: (1127-1213)

Std Err of Y Est Degrees of Freedom No. of Observations R Squared Constant Regression Output: 0.043000 0.996700 7.474100

Std Mean Maximum Minimum Terrestrial Intensity 1990.524 26.379 2032 1931

Kd Coefficient(s) Std Err of Coef. 0.246884 0.003994 R Squared
No. of Observations Std Err of Y Est

Constant

Regression Output:

Degrees of Freedom

Std Err of Coef Kd Coefficient(s)

0.249889 0.001607

Site 2, Date 910711, Time: (1334-1413) Discharge 759-770 m³/s

Location: 23.8 km

Std Err of Y Est R Squared Degrees of Freedom No. of Observations Constant Regression Output: 0.874816 0.22852 7.37234 Mean Maximum Minimum Terrestrial Intensity 1947.024 30.503 2006 1906

Kd Coefficient(s) Std Err of Coef. 0.019404 0.32034

Site 3, Date 910523

NO DATA COLLECTED

Location: 27.74 km Site 3, Date 910712, Time: (1433-1524) Discharge 142 m³/s

R Squared Std Err of Y Est Constant Degrees of No. of Obs Regression Output: 0.122103 0.969036 7.273514 55 57

Mean Maximum Minimum

> 1681.578 43.587

1773 1624

Terrestrial Intensity

Std Err of Coef. Kd Coefficient(s) 0.007791 0.323267

Site 4, Date 910713, Time: (0928-1030) Discharge 142 m³/s

Location: 50 km

Site 4, Date 910524, Time: (1101-1131) Discharge 425 m³/s

Location: 50 km

Maximum Minimum Terrestrial Intensity 1960.462 19.007 1884 1992

R Squared Std Err of Y Est Degrees of Freedom No. of Observations Constant Regression Output: 0.989854 0.101877 7.241084

ဌ ဌ Mean Maximum Minimum Terrestrial Intensity 1579.963 47.65 1505 1654

Kd Coefficient(s) Std Err of Coef.

Degrees of Freedom No. of Observations

R Squared Std Err of Y Est

0.997157

Mean Std.

0.054627 7.463962

Constant

Regression Output:

0.314592 0.001901

Kd Coefficient(s) Std Err of Coef.

0.005289 0.380370

l ocation: 77 61 km	Discharge 425 m ³ /s	Site 5, Date 910524, Time: (1417-1503)
l coation: 77 61 km	Discharge 142 m ³ /s	Site 5, Date 910713, Time: (1325-1413)

Kd Coefficient(s) Std Err of Coef.	Regression Output: Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom
0.398451 0.007785	n Output: 7.464023 0.053442 0.995439 14 12
	Terrestrial Intensity Minimum 190 Maximum 204 Mean 1944.07 Std. 45.07
	Intensity 1900 2047 1944.071 45.071
Kd Coefficient(s) Std Err of Coef.	Regressio Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom
0.367692 0.002605	n Output
	7.405623 0.101746 0.995202 98 96
	Terrestrial Intensity Minimum 1876 Maximum 1974 Mean 1928.357 Std. 27.757
	l Intensity 1876 1974 1928.357 27.757

Location: 99.5 km	Discharge 425 m³/s	Site 6, Date 910525, Time: (1009-1035)	
Location: 99.5 km	Discharge 142 m³/s	Site 6, Date 910714, Time: (0921-1005)	

Kd Coefficient(s) Std Err of Coef.	Degrees of Freedom	No. of Observations	R Squared	Std Err of Y Est	Constant	Regression Output
0.368111 0.004619	31	33	0.995141	0.054608	7.411833	n Output:
		Std.	Mean	Maximum	Minimum	Terrestrial Intensity
		41.647	1832.696	1895	1756	Intensity
Kd Coefficient(s) Std Err of Coef.	Degrees of Freedom	No. of Observations	R Squared	Std Err of Y Est	Constant	Regressio
0.399570 0.006096						egression Output:
	49	51	38721	0.079750	37072	
		Std.	Mean	_	Minimum	Terrestrial Intensity
		42.878	1497.784	1573	1430	itensity

pts
Err
9
Coef.
0.0046

Site 7, Date 910525

NO DATA COLLECTED

Site 7, Date 910714, Time: (1245–1321)
Discharge 142 m³/s

Kd Coefficient(s) 0.317502 Std Err of Coef. 0.002679	Degrees of Freedom	No. of Observations	R Squared	Std Err of Y Est	Constant	Regression Output:
ଦ ର	69	71	0.995109	0.062718	7.565349	
		Std.	Mean	Maximum	Minimum	Terrestrial Intensity
		40.206	2207.873	2285	2141	Intensity

Terrestrial Intensity
Minimum 1745
Maximum 1893
Mean 1827.843
Std. 49.812

Location: 128.56 km	Discharge 425 m³/s	Site 8, Date 910525, Time: (1510-1540)
Location: 128.56 km	Discharge 142 m³/s	Site 8, Date 910714, Time: (1415–1533)

	0.378790 0.002427	Kd Coefficient(s) Std Err of Coef.			0.506195 0.007266	Kd Coefficient(s) 0 Std Err of Coef. 0
49		Degrees of Freedom			42	Degrees of Freedom
51		No. of Observations	38.373	Std.	44	No. of Observations
0.997992		R Squared	1605.045	Mean	0.991418	R Squared
0.045507		Std Err of Y Est	1682	Maximum	0.119274	Std Err of Y Est
7.466106		Constant	1543	Minimum	7.322197	Constant
	Output:	Regression Output	ntensity	Terrestrial Intensity	uput	Regression Output

Site 9, Date 910526, Time: (1115–1143) Discharge 425 m³/s Location: 149.38 km	
Site 9, Date 910715, Time: (1024–1137) Discharge 142 m³/s Location: 149.38 km	

					FOCAGON: 149.00 KIII				
Regression Output	Output:		Terrestrial Intensity	ntensity	Regression O	n Output:		Terrestrial	Intensity
-		7.561991	Minimum	1900	Constant	•	.321053	Minimum	1751
#YEst	_	0.281489	_	2005	Std Err of Y Est	•).155348	Maximum	1900
ă	_	0.987490		1962.204	R Squared	0	0.992167	Mean	1816.786
bservations		88	Std.	23.694	No. of Observations		75	Std. 43.878	43.878
of Freedom		86			Degrees of Freedom		73		
icient(s) (0.769448 0.009338				Kd Coefficient(s) Std Err of Coef.	0.338860 0.003523			

Kd Coefficient(s) Std Err of Coef.

Constant
Std Err of Y Est

R Squared

Degrees of Freedom No. of Observations

Location: 165.47 km	Discharge 142 m ³ /s	Site 10, 910629, Time: (1011-1050)

Kd Coefficient(s) 0.742538 Std Err of Coef. 0.006669	Site 10, Date 910526, Time: (1558–1606) Discharge 425 m³/s Location: 165.47 km Regression Output: 7.21506 Std Err of Y Est 0.12605 R Squared 0.99806 No. of Observations Degrees of Freedom 2
6 6	558-1606) 7.215061 0.126055 0.998067 26 24
	Terrestrial Intensity Minimum 129 Maximum 1366.07 Mean 1366.07 Std. 17.01
	ntensity 1298 1383 1366.076 17.018
Kd Coefficient(s) Std Err of Coef.	Site 10, 910629, Time: (1011-1050) Discharge 142 m³/s Location: 165.47 km Regression Output: Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom
0.467770 0.004504	e: (1011-1050 n Output: 7
	50) 7.288689 0.133975 0.993367 74 72
	Terrestrial Intensity Minimum 172 Maximum 187 Mean 1802.37 Std. 39.25
	Intensity 1728 1877 1802.378 39.256

Site 13, Date 910527, Time: (1149-1209) Discharge 425 m³/s

Location: 200.38 km

Degrees of Freedom	No. of Observations	R Squared	Std Err of Y Est	Constant	Regression Output:
59	61	0.988256	0.270699	7.815037	
	Std.	Mean	Maximum	Minimum	Terrestrial Intensity
	9.094	2015.606	2032	2003	intensity

Kd Coefficient(s) Std Err of Coef. 0.015238 1.073789

Site 14, Date 910527, Time: (1557-1618)

Discharge 425 m³/s Location: 224.03 km

Kd Coefficient(s) Std Err of Coef. Degrees of Freedom No. of Observations R Squared Std Err of Y Est Constant Regression Output: 1.049515 0.012058 0.204789 0.994225 7.072127 46 Maximum Mean Minimum Terrestrial Intensity 1334.565

42.997

1254

1404

Site 16, Date 910528, Time: (1047–1114) Discharge 425 m³/s

Location: 249.94 km

Kd Coefficient(s) Std Err of Coef. Degrees of Freedom No. of Observations R Squared Std Err of Y Est Constant Regression Output: 1.001784 0.022662 0.973593 0.335454 7.502784

Mean Std.

1862.036

Maximum Minimum Terrestrial intensity

1819

Constant

1898

Discharge 142 m³/s Site 13, Date 910629, Time: (1525-1603)

Location: 200.38 km

R Squared Constant
Std Err of Y Est Degrees of Freedom No. of Observations Regression Output: 0.176751 7.189594 0.991040 63

> Mean Maximum

1618.523

1698

1550

40.919

Minimum

Terrestrial Intensity

Kd Coefficient(s) Std Err of Coef. 0.553968 0.006636

Site 14, Date 910630, Time: (0856-0925)

Location: 224.03 km Discharge 142 m³/s

Kd Coefficient(s) Degrees of Freedom R Squared Std Err of Y Est No. of Observations Constant Regression Output: 7.100522 0.084724 0.997251

Mean Std.

1373.79 49.689

1441 1293

Maximum Minimum

Terrestrial Intensity

Std Err of Coef. 0.542361 0.003675

Site 16, Date 910630, Time: (1206-1225)

Discharge 142 m²/s Location: 249.94 km

R Squared Std Err of Y Est No. of Observations Regression Output: 0.155124 0.9850567.692370

Kd Coefficient(s) Std Err of Coef. 0.643672 0.010889 Degrees of Freedom

Mean Maximum Minimum Terrestrial Intensity 2043.69 5.605 2052 2026

Location: 276.33 km	Discharge 425 m³/s	Site 18, Date 910528, Time: (1502-1518)
Location: 276.33 km	Discharge 142 m³/s	Site 18, Date 910630, Time: (1447-1513)

Kd Coefficient(s) Std Err of Coef.	Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom	Regression Output
1.128186 0.015108	7.306053 0.250936 0.990056 58 56	Output
	Minimum 1499 Maximum 1570 Mean 1540.37 Std. 25.07	Terrestrial
	1499 1578 1540.379 25.079	Intensity
Kd Coefficient(s) Std Err of Coef.	Constant Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom	Regression Output
0.61 3211 0.005039		
	7.337508 0.105094 0.995562 68 66	
	Minimum 167 Maximum 180 Maximum 1741.66 Mean 1741.66 Std. 32.81	Terrestrial
	1672 1808 1741.661 32.811	Intensity

Discharge 425 m³/s Location: 291.61 km	Site 20, Date 910529, Time: (1104-1122)	
Discharge 142 m³/s Location: 291.61 km	Site 20, Date 910701, Time: (0953-1023)	

Kd Coefficient(s) C	No. of Observations Degrees of Freedom	R Squared	Std Err of Y Est	Constant	Regression Output	Location: 291.61 km
0.752832 0.005484	52 50	0.997353	0.073488	7.588494	Dutput	
	Std.	Mean	Maximum	Minimum	Terrestrial Intensity	
	16.085	1864.134	1892	1837	Intensity	
Kd Coefficient(s) Std Err of Coef.	No. of Observations Degrees of Freedom	R Squared	Std Err of Y Est	Constant	Regression Output	Location: 291.61 km
0.530677 0.003500		0	0	~1	n Output:	
	62 60).997395	0.060133	.393184		
	Std.	Mean	Maximum	Minimum	Terrestria	
	40.877	1686.274	1753	1623	Terrestrial Intensity	

Discharge 425 m³/s Location: 321.06 km	Site 21, Date 910529, Time: (1529-1552)
Discharge 142 m³/s Location: 321.06 km	Site 21, Date 910701, Time: (1 425–1506)

Kd Coefficient(s) Std Err of Coef.	Regression Output Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom
1.130883 0.013118	Output
	7.319699 0.331507 0.990139 76 74
	Terrestrial Intensity Minimum 132 Maximum 147 Mean 1412.65 Std. 46.51
	ntensity 1322 1472 1412.657 46.516
Kd Coefficient(s) Std Err of Coef.	Regression Output Constant Std Err of Y Est R Squared No. of Observations Degrees of Freedom
0.599647 0.003661	n Output:
	7.346620 0.114777 0.997175 78
	Terrestrial Intensity Minimum 1728 Maximum 1873 Mean 1800,717 Std. 43.18
	Intensity 1728 1873 1800.717 43.18

Site 22, Date 910530, Time: (1047-1116) Discharge 425 m³/s Discharge 142 m³/s

Location: 350.18 km

R Squared Constant Std Err of Y Est No. of Observations Regression Output: 0.103711 0.997989 7.450815 Mean Maximum Minimum Terrestrial Intensity 1917 1873 30.794 1809

Kd Coefficient(s) Std Err of Coef. 0.933231 0.005155

Degrees of Freedom

Site 23, Date 910530, Time: (1510-1536)

Discharge 425 m³/s Location: 374.15 km

Degrees of Freedom No. of Observations R Squared Constant Std Err of Y Est Regression Output: 0.995973 0.150779 7.219155 Mean Maximum Minimum Terrestrial Intensity 1551.935 1665 1487

Std Err of Coef. Kd Coefficient(s) 1.046712 0.008591

Site 24, Date 910531, Time: (1028-1049)

Location: 387.51 km Discharge 425 m³/s

Constant Regression Output:

Kd Coefficient(s) Std Err of Coef. Degrees of Freedom No. of Observations R Squared Std Err of Y Est 1.079363 0.160836 0.994018 5.914626

Mean Std.

351.807

R Squared

430.4 283.8

Maximum Minimum Terrestrial Intensity

0.013238

Site 22, Date 910702, Time: (0852-0934)

Location: 350.18 km

Std Err of Y Est R Squared Degrees of Freedom No. of Observations Constant Regression Output: 0.123258 7.058595 0.996151

Mean Maximum 1480 1406.059 38.73

Minimum

1333

Terrestrial Intensity

Kd Coefficient(s) 0.560595 0.004321

Std Err of Coef

Discharge 142 m³/s Location: 374.15 km

Site 23, Date 910702, Time: (1314-1348)

No. of Observations R Squared Std Err of Y Est Constant Regression Output: 0.093657 0.996600 7.584349 9 6

> Mean Maximum

2029.772 17.656

Minimum

1992 2082

l errestrial intensity

Degrees of Freedom

Kd Coefficient(s)

Std Err of Coef

0.004950 0.678084

Site 24, Date 910702, Time: (1513-1541)

Location: 387.51 km Discharge 142 m³/s

Std Err of Y Est Constant Regression Output: 0.074423 7.378757

Degrees of Freedom No. of Observations 0.997813

Mean Std. Maximum Minimum Terrestrial Intensity 1632.709 48.68 1699 1550

0.690052 0.004170

Kd Coefficient(s)
Std Err of Coef.

ATTENUATION COEFFICIENT, Ka, COSINE CORRECTED

(Page 82)

Paria Tributary, Date 911106, Time: (1119-1355)
Discharge .34- .33 m³/s

Degrees of Freedom	No. of Observations	R Squared	Std Err of Y Est	Constant	Regression Output:
227	229	0.824102	1.333641	7.432745	
	Std.	Mean	Maximum	Minimum	Terrestrial Intensity
	60.11144	1258.406	1337	1143	Intensity

 Kd Coefficient(s)
 29.1333

 Std Err of Coef.
 0.89334